

## Article

# Predicted Future Changes in the Mean Seasonal Carbon Cycle Due to Climate Change

Mauro Morichetti <sup>1,\*</sup>, Elia Vangi <sup>1,2</sup> and Alessio Collalti <sup>1,3,†</sup>

<sup>1</sup> National Research Council of Italy, Forest Modelling Laboratory, Institute for Agriculture and Forestry Systems in the Mediterranean (CNR-ISAFOM), Via Madonna Alta 128, 06128 Perugia, Italy; elia.vangi@isafom.cnr.it (E.V.); alessio.collalti@cnr.it (A.C.)

<sup>2</sup> geoLAB—Laboratory of Forest Geomatics, Department of Agriculture, Food, Environment and Forestry, Università degli Studi di Firenze, Via San Bonaventura 13, 50145 Firenze, Italy

<sup>3</sup> National Biodiversity Future Center (NBFC), Piazza Marina 61, 90133 Palermo, Italy

\* Correspondence: mauro.morichetti@cnr.it; Tel.: +39-0755014541

† These authors contributed equally to this work.

**Abstract:** Through photosynthesis, forests absorb annually large amounts of atmospheric CO<sub>2</sub>. However, they also release CO<sub>2</sub> back through respiration. These two, opposite in sign, large fluxes determine how much of the carbon is stored or released back into the atmosphere. The mean seasonal cycle (MSC) is an interesting metric that associates phenology and carbon (C) partitioning/allocation analysis within forest stands. Here, we applied the 3D-CMCC-FEM model and analyzed its capability to represent the main C-fluxes, by validating the model against observed data, questioning if the sink/source mean seasonality is influenced under two scenarios of climate change, in five contrasting European forest sites. We found the model has, under current climate conditions, robust predictive abilities in estimating NEE. Model results also predict a consistent reduction in the forest's capabilities to act as a C-sink under climate change and stand-aging at all sites. Such a reduction is predicted despite the number of annual days as a C-sink in evergreen forests increasing over the years, indicating a consistent downward trend. Similarly, deciduous forests, despite maintaining a relatively stable number of C-sink days throughout the year and over the century, show a reduction in their overall annual C-sink capacity. Overall, both types of forests at all sites show a consistent reduction in their future mitigating potential.

**Keywords:** carbon cycle; climate change; process-based model; mean seasonal cycle; forest ecosystems



**Citation:** Morichetti, M.; Vangi, E.; Collalti, A. Predicted Future Changes in the Mean Seasonal Carbon Cycle Due to Climate Change. *Forests* **2024**, *15*, 1124. <https://doi.org/10.3390/f15071124>

Academic Editor: Giacomo Alessandro Gerosa

Received: 21 May 2024

Revised: 14 June 2024

Accepted: 26 June 2024

Published: 28 June 2024



**Copyright:** © 2024 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

## 1. Introduction

Forests play a pivotal role in the biosphere–atmosphere feedback by annually absorbing large amounts of atmospheric CO<sub>2</sub> through photosynthesis (GPP; ~150 PgC year<sup>−1</sup>) and releasing it back because of, e.g., ecosystem respiration (R<sub>eco</sub>), a relatively close amount yet not necessarily equal, which varies year by year [1–3]. The net ecosystem exchange (NEE) of CO<sub>2</sub> between ecosystems and the atmosphere is the net balance between these two gross fluxes opposite in sign, and it governs much of the overall terrestrial annual net carbon (C) budget. Imbalances between CO<sub>2</sub> sources (even including carbon lost by fires and other processes) and sinks directly increase or decrease atmospheric CO<sub>2</sub> levels [4]. Terrestrial ecosystems—and forests in particular—are contributing substantially to climate change mitigation, provided that they are C-sinks and not C-sources [5]. Forests that might absorb more than they emit are commonly considered carbon sinks (with NEE-negative in sign), while if they emit more than they absorb are considered as carbon sources (with NEE-positive in sign). The magnitude of this exchange of CO<sub>2</sub>, however, is subject to substantial variability and trends, in large part as a response to variations and trends in climate [6]. Indeed, forests have been shown to be extremely sensitive to changes in environmental conditions (e.g., climate, seasonality, atmospheric CO<sub>2</sub> concentration, nitrogen deposition),

to aging [7,8], and to disturbances [9], including management practices [10,11], which can control both photosynthesis and respiration. Therefore, estimating NEE, GPP, and  $R_{eco}$  is a key step for better understanding the underlying mechanisms constraining ecosystem functioning [12].

Europe and the Mediterranean are expected to become in the near future a ‘Hot Spot’ of climate change [13–16]. The literature reports that under climate change scenarios forests are expected to grow faster, to mature earlier but also to die younger, curtailing their life span [17], because of, mainly, warming and increased atmospheric  $CO_2$  concentration (the so-called ‘ $CO_2$ -fertilization effect’) [18]. Conversely, there is a general lack of evidence and knowledge on how, overall, forest ecosystems will, on the whole, react to climate change. Forest carbon balance will be impacted by climate change because various main processes are impacted, which, in turn, may react and respond differently to climate change, also because they are vulnerable and sensitive to separate environmental drivers. Indeed, while an extensive line of evidence shows that the increased availability of  $CO_2$  may amplify the photosynthetic rate and assimilation capacity [19], such an increase is largely debated since there is no evidence that such positive changes will generally continue indefinitely [20,21]. Similarly, the effects of warming are largely discussed because, while it is documented that some species may take advantage of a longer vegetation season (e.g., deciduous species), there are also negative effects linked to warming such as heat waves and the often-associated drought events [22,23], including late frost [24] and disturbances [25], which can be detrimental to growth till tree survival. In addition, there is a general concern that the changing temperature response of respiration turns boreal forests from carbon sinks into carbon sources [26]. Indeed, warming has also been found to accelerate both autotrophic as well as heterotrophic respiration (the two components of ecosystem respiration), meaning that increased temperature may lead forests to release more carbon, potentially more than absorbed annually [27]. Conversely, drought has been shown to reduce microbial respiration and then heterotrophic respiration [28]. How these processes (i.e., photosynthesis and ecosystem respiration) will be impacted by climate change annually will determine much of the future forest annual C-budget.

The Mean Seasonal Cycle (MSC) metric, which reflects the average distribution of flux (i.e., NEE, GPP, and  $R_{eco}$ ) throughout the days of a year, is an insightful measure linking phenology with carbon partitioning and allocation within seasonal climatic variability. By capturing the typical fluctuations in a specific region due to changing seasons, the MSC highlights the expected seasonal changes in climate data, averaged over many years, to smooth out anomalies and emphasize the regular, cyclical nature of these changes. Many studies [29–32] have indeed shown that climate change will impact both the phenology by changing the date for the beginning and the end of the growing season, as well as by changing the shape of the Leaf Area Index (LAI) distribution over the year, which, at the same time, will influence the way, among the other things, when photosynthesis can start and how recent and old photosynthates are partitioned and used to build new tissues and to replenish the reserves used for the metabolism, as well as carbon allocation [33] and the C-dynamic.

Process-based models are valuable tools to understand how and to what extent future climate change will impact these two fluxes (GPP and  $R_{eco}$ ) in the MSC, both processes being controlled by warming and changes in precipitation regime and atmospheric  $CO_2$  concentration [34,35]. Here, we at first applied and validated, under current observed climate conditions, the ‘Three Dimensional–Coupled Model Carbon Cycle–Forest Ecosystem Module’ (3D-CMCC-FEM), a biogeochemical, biophysical process-based forest ecosystem model designed to simulate carbon, nitrogen, and water cycle in forest ecosystems and, secondly, under climate change conditions.

Specifically, we question and analyze: (1) the capability of the 3D-CMCC-FEM to represent under the current climate the main C-fluxes governing the C-cycle in terms of net ecosystem exchange (NEE), gross primary production (GPP), and ecosystem respiration ( $R_{eco}$ ), by validating the model against independent data from the Fluxnet network; and,

provided that the model is close to the observed data; (2) how, and if, the sink/source mean seasonality will be influenced under two locally bias-corrected scenarios of warming (RCP 2.6 and 6.0) and atmospheric CO<sub>2</sub> enrichment from three CMIP5 Earth System Models, within ISIMIP-PROFOUND initiative, in five well-studied and long-monitored contrasting forest sites (three evergreens and two deciduous) on a longitudinal transect through Europe up to the end of the century.

## 2. Materials and Methods

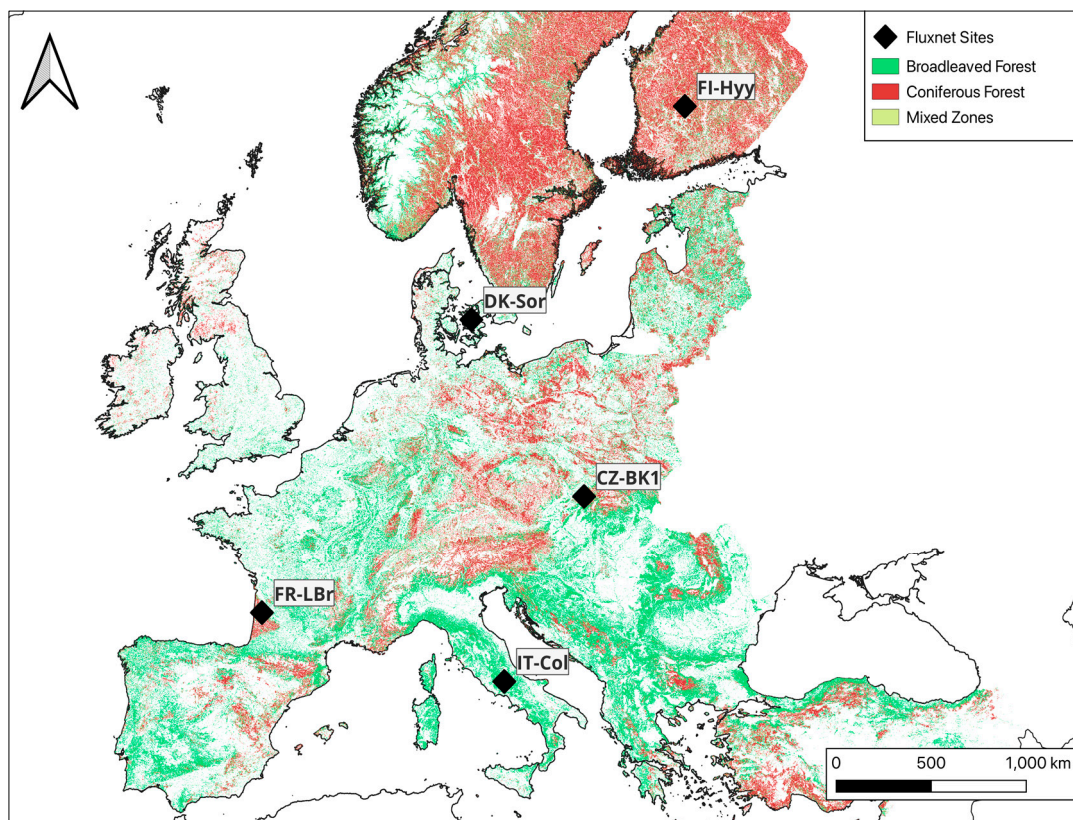
### 2.1. Model Description (3D-CMCC-FEM ‘v.5.6’)

The ‘Three Dimensional–Couple Model Carbon Cycle–Forest Ecosystem Module’ (hereafter ‘3D-CMCC-FEM’) is a biogeochemical, biophysical, process-based forest ecosystem model (see [10,11,36–44] and reference therein). The model is designed to simulate carbon, nitrogen, and water cycles in forest ecosystems at commonly 1-hectare spatial resolution and the main eco-physiological processes (e.g., photosynthesis) at daily temporal resolution. The most recent code versions since Collalti et al. [17] adopt the biogeochemical photosynthesis model of Farquhar, von Caemmerer, and Berry [45] to compute gross primary productivity (GPP). The biogeochemical photosynthesis model is parameterized as in Bernacchi et al. [46,47] and temperature acclimation for leaves as in Kattge and Knorr [48]. The 3D-CMCC-FEM considers, as in De Pury and Farquhar [49], light interception, reflection, transmission, and assimilation (and leaf respiration) for both sun and shaded leaves. Autotrophic respiration ( $R_A$ ) is computed mechanistically following the ‘Growth and Maintenance Respiration Paradigm’ (GMRP; [50]), which is divided into the metabolic costs for synthesizing new tissues (growth respiration,  $R_G$ ) and the metabolic costs for maintaining the existing ones (maintenance respiration,  $R_M$ ). In 3-D-CMCC-FEM, the maintenance respiration is based on Nitrogen amount (a fixed fraction of carbon mass varying between the six tree compartments) and is temperature-controlled by a standard Arrhenius relationship [36]. ‘Type I’ and ‘Type II’ acclimation of respiration to temperature (i.e., short- and long-term acclimation; [10,39,51]) are also accounted for. Any imbalance between carbon assimilation and carbon losses because of plants’ respiration is buffered by a seventh pool, the Non-Structural Carbon pool (NSC; starch and sugars undistinguished), which has priority in the carbon allocation all over the year. The net primary production (NPP) is the GPP minus  $R_A$ . Biomass production (BP) is the amount of NPP not used for replenishing the NSC pool. Indeed, other forms of non-structural carbon losses (e.g., biogenic volatile organic compounds, BVOCs, or root exudates to mycorrhizas) are not accounted for by the model. The phenological scheme, as well as the carbon partitioning/allocation scheme, distinguished for deciduous and evergreen tree species, is in-depth described in Collalti et al. [10,36,39,52]. Heterotrophic respiration follows a BIOME-BGC-like approach (which follows the CENTURY model; [53,54]) distinguishing decomposition for litter and soil pools with each of the four different conceptual pools characterized by different decomposability degrees (i.e., fast, medium, slow, and a recalcitrant carbon pool) [39,55]. Altogether, litter and soil decomposition emissions form heterotrophic respiration ( $R_H$ ), which, summed up to  $R_A$ , constitutes ecosystem respiration ( $R_{eco}$ ). Net ecosystem exchange (NEE) is equal to  $R_{eco}$  minus GPP. Therefore, negative values indicate carbon uptake from the atmosphere (i.e., the system acts as a C-sink,  $NEE < 0$ ), and positive values indicate carbon release (i.e., the system acts as a C-source,  $NEE > 0$ ). The 3D-CMCC-FEM’s sensitivity to its 54 species-specific parameters and how it varies along the forest development and under climate change is described in Collalti et al. [17].

In the present study, we used version 5.6 [41], which slightly differs from v.5.5, described in Collalti et al. [8] and Dalmonech et al. [11], for a new scheme (and relative parameterization) for sapwood and live wood turnover and dynamics and some additional new forest management schemes (not used here) described in Testolin et al. [43].

## 2.2. Case Study Areas

Five case studies were selected as representative of the main European forest species (and climate) and at the same time because of long-monitored sites and part of the Fluxnet network [56], the ISIMIP (Inter-Sectoral Impact Model Intercomparison Project, [57]; <https://www.isimip.org/>, accessed on 1 January 2020) initiative and the PROFOUND database [42,58,59]: the temperate European beech (*Fagus sylvatica* L.) forest of Collelongo, Italy (IT-Col), and of Sorø, Denmark (DK-Sor), the maritime pine (*Pinus pinaster* Ait.) forest of Le Bray, France (FR-Lbr), the boreal Scots pine (*Pinus sylvestris* L.) forest of Hyytiälä, Finland (FI-Hyy), and the temperate Norway spruce (*Picea abies* (L.) H. Karst) forest of Bílý Kříž, Czech Republic (CZ-Bk1) (Figure 1). Stand characteristics are described in Table 1.



**Figure 1.** Forest classification for broadleaved and coniferous forests at the European level in the spatial resolution of 100 m [60]. Location of the Fluxnet sites considered.



**Table 1.** Site description and stand initialization data are used in simulations with data from the PROFOUND database [58]. Model initialization data (e.g., diameter at breast height—DBH, tree height, stand age, and stand density) correspond to the stand conditions of the first year in the historical simulations.

Fluxnet Code	Nation	Site Name	Coordinate			Climate	Species Composition	Period of Simulation and Climate Data		Mean DBH (cm)	Tree Height (m)	Mean Stand Age (years)	Stand Density (n. trees ha <sup>−1</sup> )
			Lat (°)	Lon (°)	Elevation (m)			Historical (Observed)	Historical + RCP (Modeled)				
IT-Col	Italy	Collelongo	41.8	13.6	1560	Temperate	<i>Fagus sylvatica</i>	1997–2014	1997–2099	20.2	17.3	105	900
FR-LBr	France	Le Bray	44.7	−0.8	61	Temperate	<i>Pinus pinaster</i>	1996–2008	1997–2099	26.7	17.8	26	614
FI-Hyy	Finland	Hyytiälä	61.9	24.3	181	Boreal	<i>Pinus sylvestris</i>	1996–2014	1997–2099	13	11.3	35	870
CZ-Bk1	Czech Republic	Bílý Kříž	49.3	18.3	875	Temperate	<i>Picea abies</i>	2000–2008	1997–2099	9.1	7.5	19	2388
DK-Sor	Denmark	Sorø	55.3	11.4	40	Boreal	<i>Pinus sylvestris</i>	1996–2008	1997–2099	24.3	21	75	353

### 2.3. Input, Meteorological Data, and Climate Change Scenarios

To run, 3D-CMCC-FEM needs a set of input data from state variables representing the stand at the beginning of simulations and that account for structural characteristics (e.g., tree height, average age, diameter at breast height; see Table 1) as well as carbon and nitrogen pools (e.g., stem carbon and nitrogen); meteorological forcing data (e.g., daily maximum and minimum temperature, daily precipitation); annual atmospheric CO<sub>2</sub> concentrations; and species-specific parameters (e.g., maximum stomatal conductance). All the data used to initialize the model in the present study for the five stands come from the ISIMIP initiative and the PROFOUND database [58]. More specifically, daily observed meteorological data for model validation come from the Fluxnet2015 Dataset [56], while daily modeled historical (1997–2005) and future climate scenarios (2006–2099) are those from the ‘ISIMIP 2bLBC’ experiments (‘2b experiments Locally Bias Corrected’) coming from three different Earth System Models (ESMs; GFDL-ESM2M, IPSL-CM5A-LR, and MIROC5, respectively) based on the Climate Model Intercomparison Project 5 (CMIP5) driven by two Representative Concentration Pathways (RCPs) of atmospheric greenhouse gas concentration trajectories, namely, RCP 2.6 and RCP 6.0. [61,62]. The ISIMIP 2bLBC have the same structure as those in the 2b experiments but have been corrected by improving the method described in Hempel et al. [63] and subsequently by the methods described in Frieler et al. [57] and Lange [64] using the observed meteorology at the local level [58]. Therefore, the 2bLBC climate data represent the more consistent and closer modeled climate data with the observational data. The annual atmospheric CO<sub>2</sub> concentrations for the historical period are based on Meinshausen et al. [65] and have been extended for the period from 2006 to 2015 with data from Dlugokencky and Tans [66]. Values specific for each RCP for the period 2016 to 2099 are also based on Meinshausen et al. [65] and were used within the Farquhar, von Caemmerer, and Berry [45] photosynthesis model with values varying, at the end of this century, from 421.4  $\mu\text{mol mol}^{-1}$  (RCP 2.6) to 666.4  $\mu\text{mol mol}^{-1}$  (RCP 6.0), respectively. NEE, GPP, and R<sub>eco</sub> data, for model validation, come from the Fluxnet2015 Dataset [56]. Other variables have been validated at these forest stands in the past (although using slightly different model versions) and described in Collalti et al. [8,10,37], Marconi et al. [40], Mahnken et al. [42], and Dalmonech et al. [11].

### 2.4. Model Runs, Validation, and Mean Seasonal Cycle Analyses under Climate Change

The model simulations for model validation under measured forcing climate ran for CZ-Bk1 from 2000 to 2008, for IT-Col from 1997 to 2014, for FI-Hyy from 1996 to 2014, for FR-Lbr from 1996 to 2008, and for DK-Sor from 1996 to 2008 (see Table 1). For all sites model simulations under climate change scenarios began in 1997 and finished in December 2099. Model validation was performed by comparing modeled NEE, GPP, and R<sub>eco</sub> against measured eddy covariance estimates (for GPP and R<sub>eco</sub> using the night-time method with constant USTAR [67], as reported in the Fluxnet2015 Dataset [56].

To analyze 3D-CMCC-FEM’s capabilities to simulate NEE, GPP, and R<sub>eco</sub> for daily and monthly time series, a set of commonly used statistical metrics have been applied to compare measured vs. modeled data (under observed climate forcing). To avoid considering bad quality data, a filtering procedure for quality-check has been applied; days with less than 60% of valid data were not considered and excluded both in the model and in the observed datasets. Therefore, daily NEE, GPP, and R<sub>eco</sub> eddy covariance data with low-quality values (i.e., less than 0.6; [68]) were removed. Consequently, the corresponding daily modeled data were removed too. The monthly NEE, GPP, and R<sub>eco</sub> values (both from eddy covariance and the model) were aggregated from the daily ones. The common statistic we applied includes Pearson’s correlation coefficient (*r*), Relative Mean Bias (RMB), Normalized Root Mean Square Error (NRMSE), and Modeling Efficiency (ME)

In climate change projections, we considered the potential modifications in the ability of forest stands to absorb or emit carbon throughout the season and across the years under two different locally bias-corrected climate change scenarios, each coming from three Earth System Models. This involves averaging the daily values of the MSC of the three

fluxes considered. It is noteworthy that NEE represents the equilibrium between carbon absorption by vegetation during photosynthesis and carbon release through vegetation and microbial respiration. Not only the length of the growing season but also the balance between the yearly amount of photosynthesis and  $R_{eco}$  has been shown to control much of the variability across the sites and the decades analyzed [47,49]. This calculation is derived from the variance between GPP and  $R_{eco}$  encompassing both autotrophic respiration ( $R_A$ , including ground components) and heterotrophic respiration ( $R_H$ ) [69].

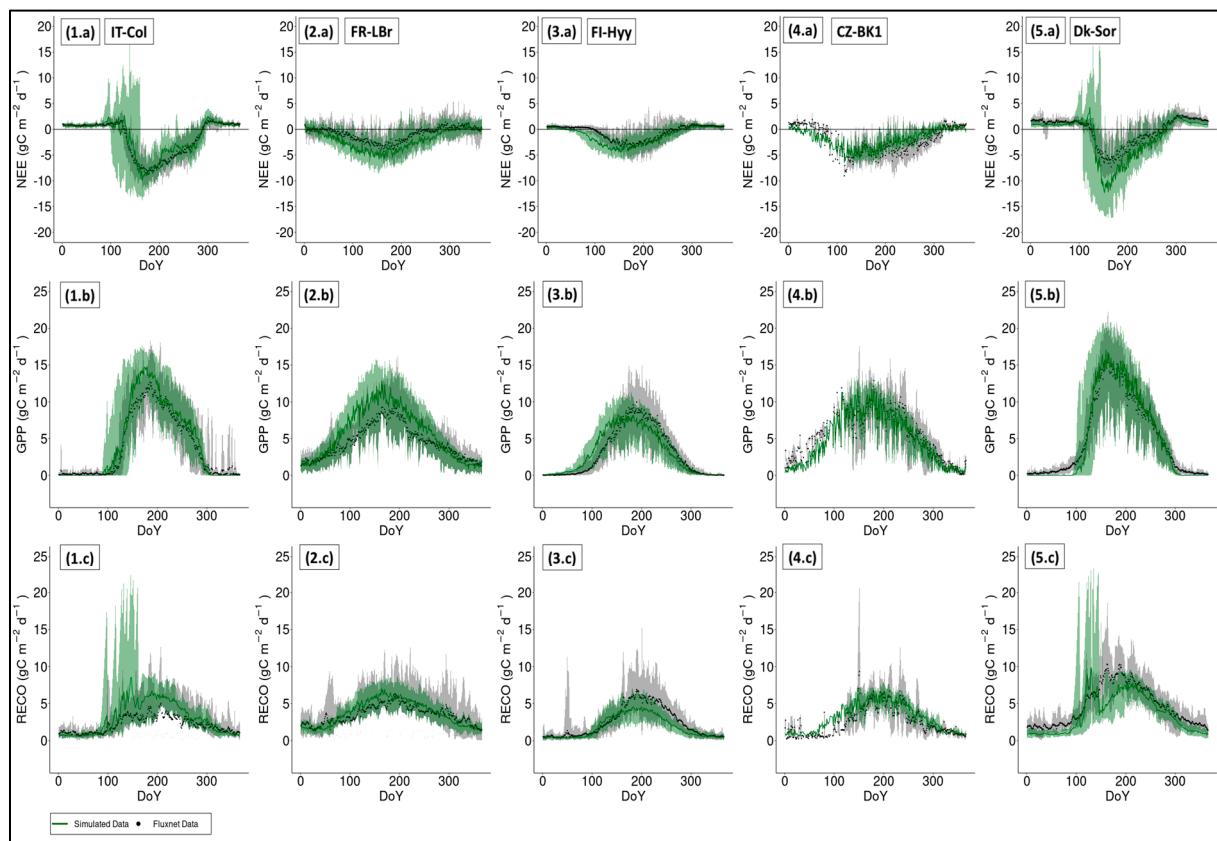
Under climate change scenarios, we account for potential changes in the sink/source capacity of the stands during the season by averaging NEE, GPP, and  $R_{eco}$  values every ten years up to 2100 [70] and accounting for changes in the sink/source and source/sink length during the year, computed as the number of total days of the year (DoY) where a forest stand behaves as a C-sink ( $GPP > R_{eco}$  with  $NEE < 0$ ) or a C-source ( $GPP < R_{eco}$  with  $NEE > 0$ ), as described by the NEE. In addition, we also account for the changes in the DoY where NEE changes its sign at least for ten consecutive days to avoid artifact effects of pulsing, e.g., the ‘Birch effect’ on  $R_{eco}$  [71], and to account for unstable conditions and no clear source/sink and sink/source seasonal transition during the year. Therefore, we discuss changes in the MSC under the RCP 2.6 and RCP 6.0 for NEE, being the net result of opposite fluxes (i.e., GPP and  $R_{eco}$ ), through its changes in negative values, i.e., days in the year where  $NEE < 0$  and describing  $CO_2$  uptake from the atmosphere, and positive, i.e., days in the year where  $NEE > 0$  and describing  $CO_2$  release to the atmosphere. In this way, we account for potential changes that may lead to anticipations or delays in the switch from source/sink and sink/source capacity, which often happens during spring and autumn during the year. The analyses under climate change scenarios (2006–2099) also include the changes in the shape of the Mean Seasonal Cycle (MSC) for NEE, GPP, and  $R_{eco}$  values and for the changes (both in the absolute and the percentage values) in the annual value. Changes in MSC were estimated on the ten-year average values using the 1998–2008 decade as a benchmarking reference. Furthermore, we account for changes in the annual values of NEE, GPP, and  $R_{eco}$  due to climate change.

### 3. Results

#### 3.1. Model Validation vs. Fluxnet Data

The NEE, GPP, and  $R_{eco}$ , as modeled by 3D-CMCC-FEM, exhibit strong correlations with the observed daily and monthly eddy covariance data and for the overall MSC across all five sites (Figures 2 and A1 and Tables A1–A3 in Appendix A). Some slight differences are observed for the daily values for temperate European beech forests of IT-Col and DK-Sor when representing the NEE and  $R_{eco}$  between the 100 and 200 DoY (Figure 2(1.a–1.c) and Figure 2(5.a–5.c)). Modeled overestimations for GPP of about  $5 \text{ gC m}^{-2} \text{ day}^{-1}$  at the peak of production (~180 DoY) for both the European beech forest at IT-Col (Figure 2(1.b)) and the maritime pine forest at FR-Lbr (Figure 2(2.b)) was observed. The highest correlation coefficients between modeled and observed data were observed for DK-Sor and IT-Col ( $r = 0.97$  and  $0.96$ , respectively) for the daily NEE and GPP (Table A1, Appendix A), while FI-Hyy shows the best correlation for daily  $R_{eco}$  ( $r = 0.94$ ) and monthly NEE ( $r = 0.99$ ) (Tables A2 and A3).

At the IT-Col and CZ-Bk1, the model shows the best performances in simulating NEE (RMB = 0.07 and 0.14, respectively—Table A1), while the lowest values were reached by DK-Sor and IT-Col for the GPP (RMB ranging between  $-0.06$  and  $0.55$ , Table A2). Overall, RMB values for GPP are relatively low across all case studies and time scales. Regarding  $R_{eco}$  at FI-Hyy and DK-Sor, the model tended to underestimate both daily (RMB of  $-0.3$  and  $-0.42$ , respectively) and monthly ( $-0.29$  and  $-0.42$ , respectively). Conversely, at IT-Col, the model exhibited a slight overestimation (RMB daily  $0.82$  and monthly  $0.77$ , Table A3).



**Figure 2.** The green lines represent the modeled (a) NEE, (b) GPP, and (c)  $R_{eco}$  amounts ( $\text{gC m}^{-2} \text{ day}^{-1}$ ) per DoY (Day of Year) for the five selected case studies, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-Lbr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor compared to relative observed data (depicted as black dots) from the Fluxnet2015 Dataset [56]. The lower and upper lines of the shaded area represent, respectively, the minimum and maximum values of the observed and modeled datasets considered.

The model reports negative NRMSE values for NEE across all time scales, indicating a slight overestimation (values ranging from  $-0.57$  to  $-6.59$ ) with the exception of the CZ-Bk1 site for the daily NEE value (1.39), and at the IT-Col site for the monthly NEE value (0.99). Regarding the GPP, (Table A2) DK-Sor showed the lowest NRMSE for daily and monthly values (0.19 and 0.16, respectively). Conversely, at CZ-Bk1 and FR-Lbr forests, the model displays the highest NRMSE for daily (1.21) and monthly (0.40) values, respectively. Last, Table A3 shows the validation results for  $R_{eco}$ . At FR-LBR, the highest accuracy and precision were reported with the lowest NRMSE for daily and monthly values (0.24 and 0.20 respectively). For the other sites, we found almost the same model capability described for the above fluxes (i.e., NEE and GPP) with a lower correlation and a slightly higher error in terms of daily  $R_{eco}$  at CZ-Bk1 (NRMSE = 1.16).

ME exhibits values close to one across all time scales and sites for NEE, with values ranging from  $-0.10$  (FR-LBr) to  $0.92$  (IT-Col) for daily NEE values and  $-0.15$  (FR-LBr) to  $0.95$  (IT-Col) for monthly values. The modeled GPP has a similar trend of the observed NEE with values ranging from  $0.28$  for FR-LBr to  $0.97$  for DK-Sor for daily values and FR-LBr =  $0.34$  to DK-Sor =  $0.92$  for monthly values (Table A2). The  $R_{eco}$  achieved lower performances than GPP and NEE in terms of ME; Table A3 displays values from  $1.17$  for IT-Col to  $0.75$  for FI-Hyy and IT-Col =  $-0.64$  to FI-Hyy =  $0.77$ , for daily and monthly time scale, respectively.

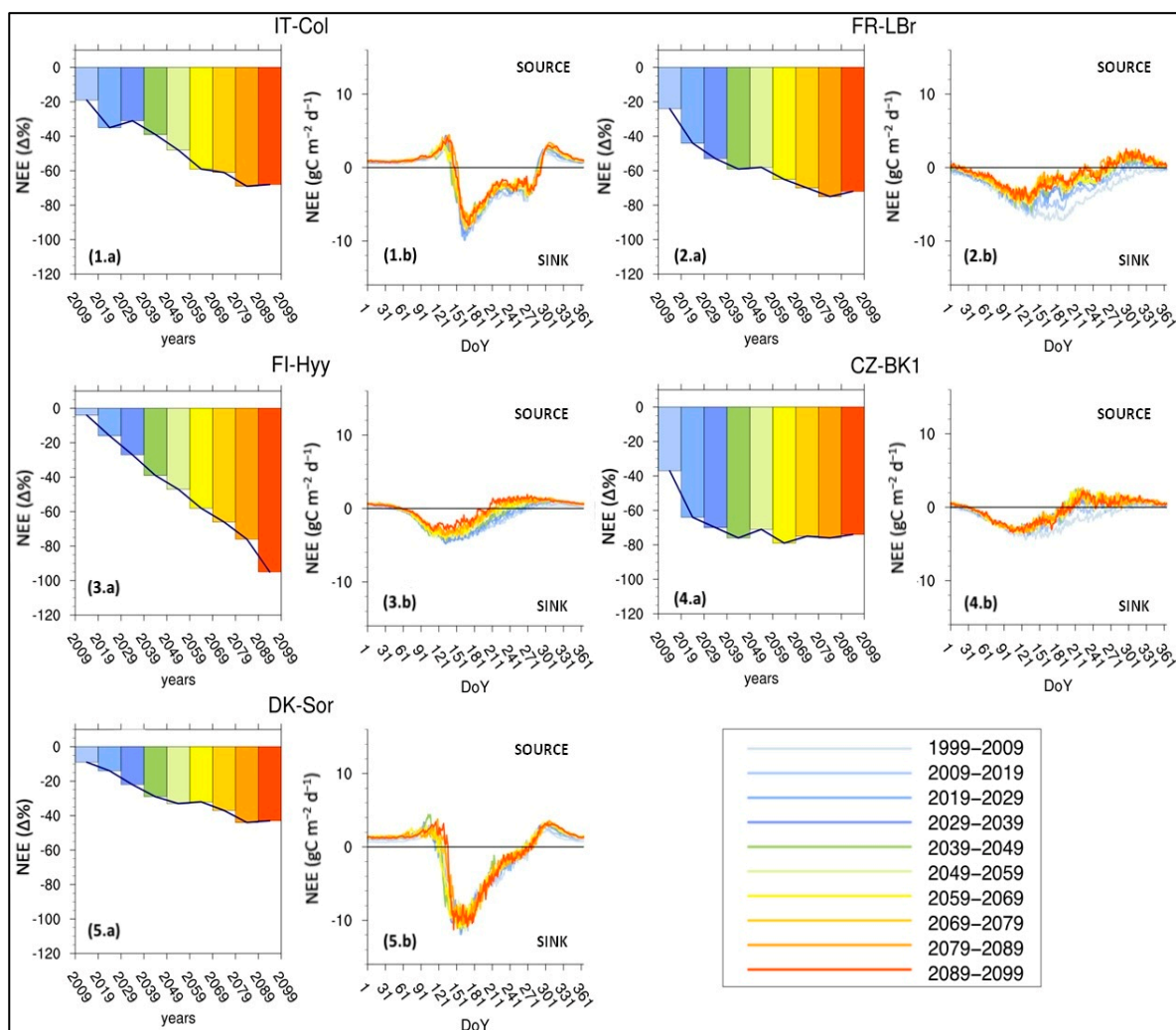
Finally, the lowest MAE was found at IT-Col, for NEE ( $0.67 \text{ gC m}^{-2} \text{ day}^{-1}$  and  $11.42 \text{ gC m}^{-2} \text{ month}^{-1}$ ) and GPP ( $0.91 \text{ gC m}^{-2} \text{ day}^{-1}$  and  $20.68 \text{ gC m}^{-2} \text{ month}^{-1}$ ), while



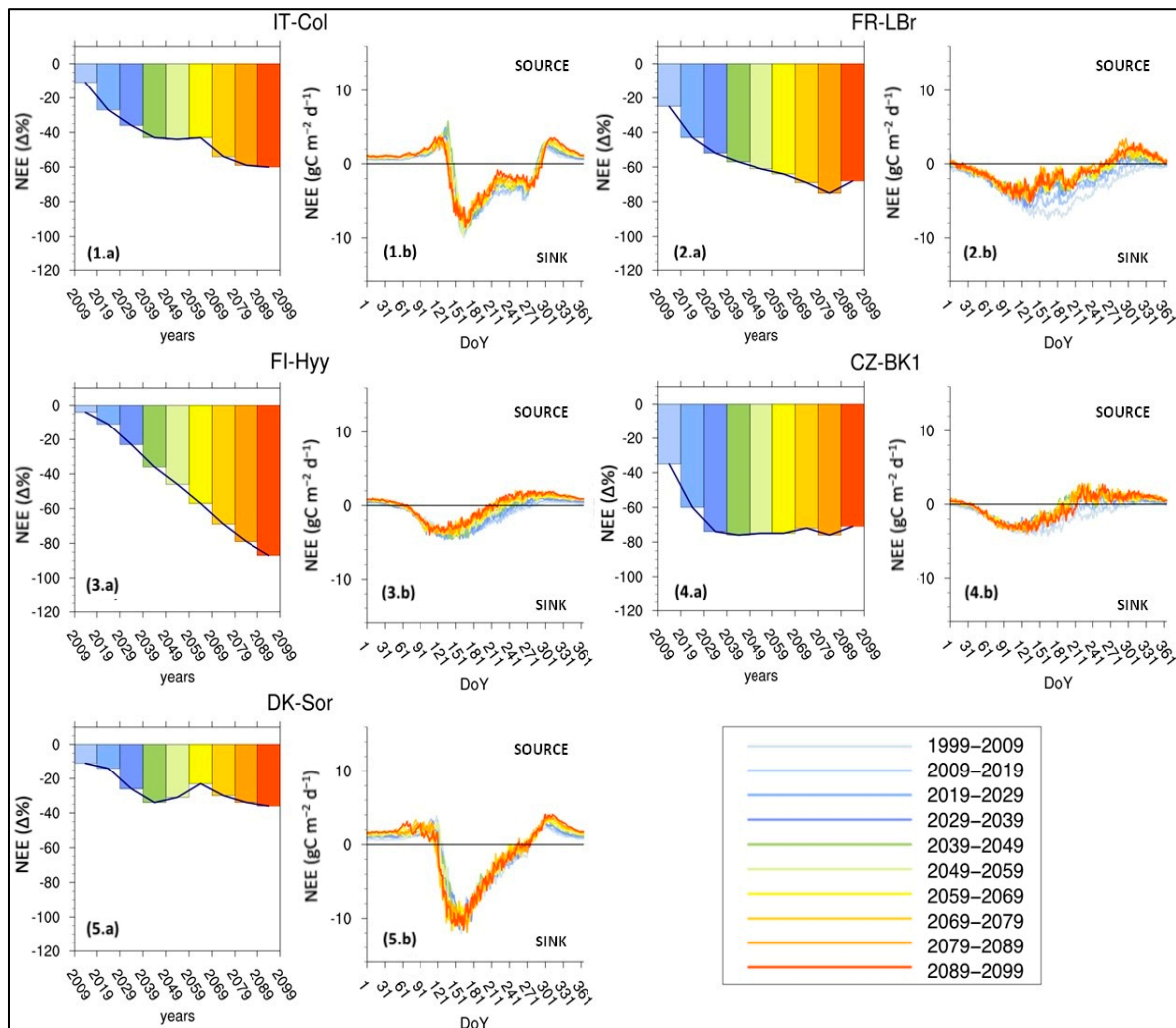
the lowest MAE for  $R_{eco}$  was for FR-LBr ( $0.71 \text{ gC m}^{-2} \text{ day}^{-1}$  and  $18.01 \text{ gC m}^{-2} \text{ month}^{-1}$ ) (Tables A1–A3 in Appendix A).

### 3.2. Mean Seasonal NEE Cycle under Climate Change Scenario

Figures 3 and 4 display the 10-year average NEE seasonal cycle under climate scenarios (i.e., RCP 2.6 and RCP 6.0) for the five case studies. Overall, across all the sites and scenarios considered, there is a consistent reduction in the absolute NEE over time (i.e., NEE is ‘less negative’ showing a reduction in the sink capacity) with changes in the source/sink (NEE becomes negative and the site turns into a C-sink) and sink/source (NEE becomes positive and the site turns into a C-source) switch over the year(s). However, RCP 2.6 generally exhibits lower reductions in annual and mean seasonal NEE when compared to RCP 6.0 across most study sites and time intervals. The loss in the modeled sink capacity is because  $R_{eco}$  increases more than GPP, and the differences between the scenarios are related to an increase in  $R_{eco}$  higher than that for GPP under RCP 6.0.



**Figure 3.** Ten-year average NEE seasonal cycle under the RCP 2.6 climate scenario for 5 case studies selected, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. The histograms (a) represent the annual NEE variation (%) from the first decade taken as a benchmark of simulation (1999–2009). The xy plots (b) show the Mean Seasonal NEE Cycle of daily values ( $\text{gC m}^{-2} \text{ day}^{-1}$ ).



**Figure 4.** Ten-year average NEE seasonal cycle under the RCP 6.0 climate scenario for 5 case studies selected, i.e., (1) Collelengo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. The histograms (a) represent the annual NEE variation (%) from the first decade taken as a benchmark of simulation (1999–2009). The xy plots (b) show the Mean Seasonal NEE Cycle of daily values ( $\text{gC m}^{-2} \text{ day}^{-1}$ ).

The rate of decrease in the annual NEE varies among locations and forest species. For example, the beech forests at IT-Col and DK-Sor show a reduction in the NEE (i.e., the site is less C-sink) of about 68%, and 43% (Figures 3 and 4—panel 1 and panel 5) at the end of the century, generally exhibiting a more moderate decrease compared to evergreen sites. The Scots pine forest shows, at the end of the century, a reduction in the sink capacity of 95%, standing out as the most significant decrease in NEE (Figures 3 and 4—panel 3). Over time, the rate of decline shows a tendency to speed up, hinting at a reduction in the capacity of the carbon sink because of, despite an overall increase in the GPP (Figures A2 and A3), the increased ecosystem respiration due to increased temperatures (Figures A4 and A5). By the end of the century, substantial reductions in NEE, i.e., the sites become less C-sink, across all locations and scenarios were simulated, with some locations experiencing over 70–90% reduction compared to the 1999–2009 decade (Table A4).

The GPP, similar to NEE, generally increases across forest types and scenarios over time, albeit with varying degrees (Figures A2 and A3). Specifically, RCP 6.0 shows higher growth rates, particularly in later years (2059–2099), with increases ranging from 28 to 51% across all studied forests, compared to 10–26% for RCP 2.6 (Table A5). Analysis of long-term trends suggests saturation and subsequent slight decreases in GPP growth

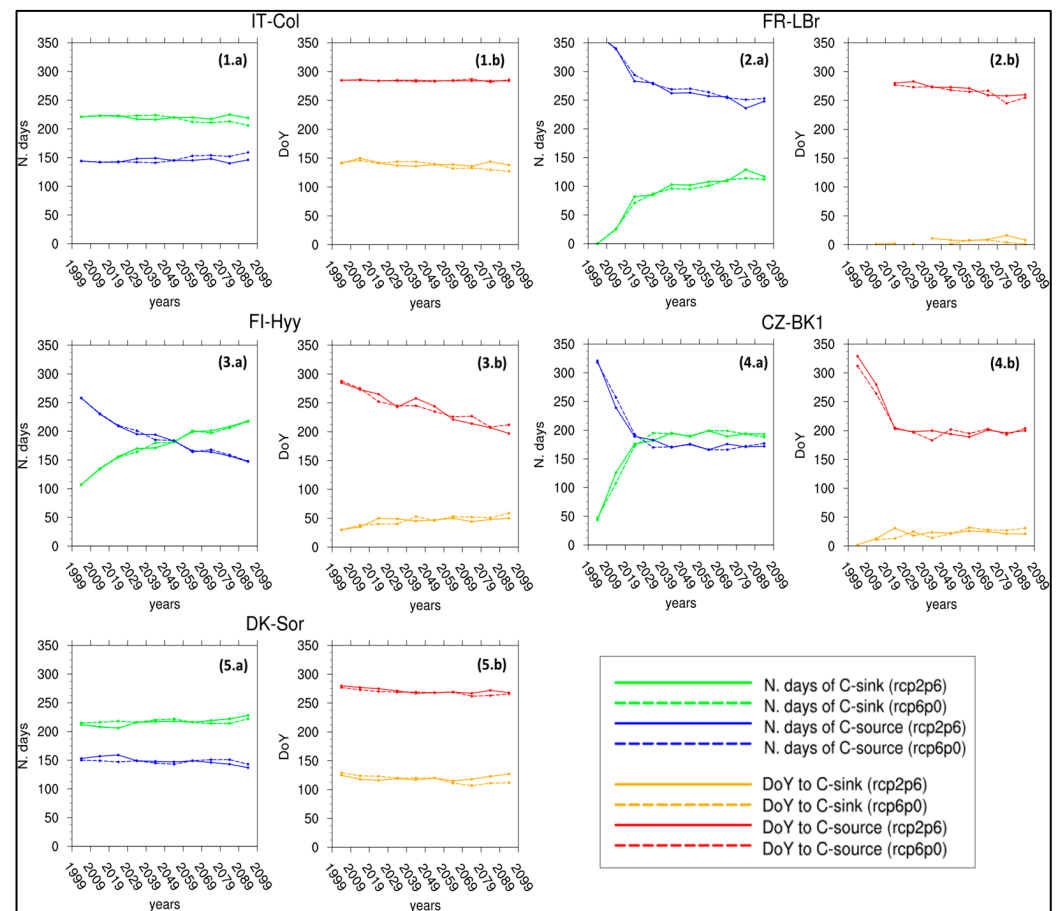
rates under RCP 2.6, with this phenomenon being most noticeable in the forests of FI-Hyy. RCP 6.0 consistently shows a higher percentage of change in  $R_{eco}$  compared to RCP 2.6. By the end of the century (2089–2099),  $R_{eco}$  changes from 37 to 106% for RCP 2.6 and 60 to 121% for RCP 6.0 across all forests examined (Table A6). Boreal and maritime pine forests experienced higher changes compared to deciduous forests. For example, by 2089–2099, under RCP 6.0, changes reach 142% and 121% at FI-Hyy and FR-Lbr forests, while deciduous forests like European beech in IT-Col and DK-Sor experience lower changes at 67% and 60% respectively.

### 3.3. Changes in NEE Dynamics under Different Climate Scenarios

We considered fluctuations in the length of sink/source and source/sink forest stand behaviors throughout the year. This calculation involved determining the total number of days in a year ('N. days year<sup>-1</sup>') where a forest stand exhibited either C-sink ( $NEE < 0$ ) or C-source behavior ( $NEE > 0$ ). The 10-year average number of days as a C-sink and as a C-source under the RCP 2.6 and 6.0 for the five case studies selected is presented in Figure 5a. The number of days identified as a C-sink in the evergreen forests for the scenario RCP 2.6 (i.e., FR-LBr, CZ-Bk1, and FI-Hyy) starts relatively low in the first decade (1999–2009) but increases significantly over time, showing a consistent upward trend (Figure 5(2.a,3.a,4.a)). At the CZ-Bk1 site, the number of days considered as a C-sink starts at 47 N. days year<sup>-1</sup> in the 1999–2009 decade and increases steadily over the decades, peaking at 199 N. days year<sup>-1</sup> in 2059–2069 and remaining relatively high thereafter. For FI-Hyy forests, the number of days as a C-sink starts at 107 N. days year<sup>-1</sup> in 1999–2009 and rises to 217 days in 2089–2099, with a peak of 225 in 2079–2089. Lastly, at the forests of FR-LBr, there were no days as a C-sink in 1999–2009, but they increased steadily over time, reaching 114 days by 2079–2089. The number of days as a C-source varies inversely to those as a C-sink across decades, demonstrating a general reduction over time for evergreen forests. The number of days functioning as a C-source decreased, from ~300 in 1999–2009 to ~200 N. days year<sup>-1</sup> in 2089–2099 (Table A1), for evergreen forests. Deciduous forests on the contrary almost maintain a relatively stable number of days as a C-sink through the century (Figure 5(1.a,5.a)). The number of days as a C-sink at IT-Col forests ranges from 212 to 224 N. days year<sup>-1</sup>, while at DK-Sor from 214 to 222 N. days year<sup>-1</sup>, with slight fluctuations observed across decades and no clear overall trend. The same trend for the capacity of the stand to act as a C-source, but with slightly different days in the range from 137 to 159 N. days year<sup>-1</sup> recorded in the forest of Sorø, and from 140 to 149 N. days year<sup>-1</sup> for the IT-Col site. The discrepancies between the already described RCP 2.6 and RCP 6.0 scenarios are minimal as they exhibit a similar trend for evergreen and deciduous, respectively, with only a slight change in the number of days, mainly in the last decade, with a magnitude of ~10 N. days year<sup>-1</sup>.

To assess the changes in MSC due to climate change scenarios, we focused on the shifting patterns of the DoY over decades, particularly examining when forest stands transitioned from a C-source to a C-sink and vice versa (see Figure 5b). Regardless of the scenario considered, the European beech forests revealed constant transition periods across decades. In IT-Col, the shift from being a C-source to a C-sink occurred between DoY 136 and 150, with the opposite transition from sink to source around DoY 285. Meanwhile, in DK-Sor, the same transitions happened between DoY 115 and 127, and reversed between DoY 272 and 280 (Figure 5(1.b,5.b)). Even for evergreen forests, there are no noticeable differences in the shift corresponding to RCP 2.6 and RCP 6.0 scenarios. At CK-BZ1, the shift to a C-sink occurs relatively early in the year, spanning from DoY 2 to 31 across the decades. At FI-Hyy, the transition timings vary widely, ranging from DoY 30 to 50 over the decades. Similarly, at FR-LBr, the C-sink transitions occur from DoY 1 to 16 across the decades, with some decades exhibiting earlier shifts. On the other hand, transitions to C-sources at CZ-BK1 occur from late November to early December (DoY 201–329), displaying a decreasing trend over the years. In contrast, at FI-Hyy, transitions take place

from mid-September to late September (DoY 197–285), while at FR-LBr, they occur from late September to early October (DoY 255–285) (Figure 5(2.a,3.a,4.a)).



**Figure 5.** Ten-year average number of days ( $N. \text{ days year}^{-1}$ ) as C-sink and as C-source (a) and DoY (Day of Year) in which C-source switch to C-sink and C-source to C-sink (b) under the RCP 2.6 and 6.0 for the 5 case studies selected, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. Data missing for some intervals are because of filtering and data removals to avoid pulsing artifacts, e.g., the ‘Birch effect’ and unstable conditions (see Section 2).

## 4. Discussion

### 4.1. Model Validation

To sum up statistical metrics, the 3D-CMCC-FEM performs best in replicating the mean seasonal patterns of the three fluxes in European beech forests DK-Sor and IT-Col.  $R_{\text{eco}}$  also reaches a satisfactory performance for the boreal Scots pine forest of FI-Hyy, and the Scots pine forest of CZ-Bk1. The robust predictive ability of 3D-CMCC-FEM in estimating NEE across different timeframes, forest species, and climates, as proved by its alignment with the eddy covariance Fluxnet2015 Dataset [56], underscores its effectiveness in capturing the complex dynamics of carbon fluxes within forest ecosystems, as documented by previous works (see, e.g., [10,11,36,40,42]). Nevertheless, there are slight inconsistencies, particularly during peak photosynthesis periods (around the 200th day of the year) in both deciduous and evergreen forests, which have been acknowledged in the existing literature. Studies indicate that estimates of ecosystem respiration derived from eddy covariance often underestimate actual values for forest ecosystems [72–75].



#### 4.2. Mean Seasonal NEE Cycle under Climate Change Scenario

The sensitivity of forest ecosystems to changes in environmental factors such as climate change, seasonal variations, and atmospheric CO<sub>2</sub> levels has been thoroughly evidenced in the literature. Indeed, the interaction among these variables plays a crucial role in shaping the carbon exchange dynamics [6,12,17,31,76].

The analysis of the mean NEE seasonal cycle under different climate change scenarios presents several fascinating findings. The first key observation is the consistent reduction in forest stand capabilities to act as carbon sinks from the atmosphere in the coming years, across all study sites and climate change scenarios. Despite the overall lack of understanding regarding how forest ecosystems will respond to climate change, several new studies concur regarding the diminishing forest carbon uptake capabilities, thus confirming our statement [26,31,77–79].

In broader terms, climate change affects forest carbon balance by influencing key processes, which can respond differently due to their sensitivity to various environmental drivers [80]. Indeed, the reduction in NEE, within the forests observed in this study, has a disparity in the intensity of climate impacts across these scenarios: RCP 2.6 exhibits less intense NEE reduction compared to RCP 6.0 scenarios. The warming accelerates both autotrophic and heterotrophic respiration meaning that increased temperature may lead forests to an increase in  $R_{eco}$  [27,58,81,82]. On the other hand, the increasing atmospheric CO<sub>2</sub> concentration intensifies the GPP through the ‘carbon fertilization effect’ (i.e., reported to be the cause of 44% of the GPP increase since the 2000s) [18,19]. Under the RCP 2.6 scenario,  $R_{eco}$  exhibits a steady linear increase until the century’s end, while GPP follows a bell-shaped curve, reaching saturation around the mid-century. Consequently, as GPP saturates, its compensatory capacity reduces, while  $R_{eco}$  continues to rise due to further temperature increases in the latter half of the century resulting in a further decline in NEE. Conversely, for the RCP 6.0 scenario, GPP saturation occurs only towards the end of the century, and while one might anticipate a reversal in NEE trends considering this factor, it does not materialize. This is because, under the RCP 6.0 scenario, a higher temperature increase is predicted compared to RCP 2.6 (i.e., [58]), leading to a more pronounced rise in ecosystem respiration relative to photosynthesis. As a consequence, there is a greater decrease in the stand forest’s carbon sink capacity. The result refers to all forests studied in this work, but in the forest of FI-Hyy, the phenomena look more pronounced (Figures A2 and A3 in Appendix A).

The fate of ecosystem carbon flux depends not only on atmospheric and climate conditions but also on the type of forests analyzed. The reduction in C-sink capabilities is particularly notable in evergreen forests, which exhibit a higher decrease in NEE compared to evergreen sites. The boreal Scots pine forest of FI-Hyy stands out with the most significant reduction in NEE, indicating a heightened vulnerability to climate change effects. This finding is supported by the study of Hadden and Grelle [26] who found that, over a 17-year period, the forest ecosystem in a boreal forest stand in northern Sweden transitioned from being a carbon sink to a carbon source. This could mean that past efforts to validate the neutrality hypothesis [83] with climate change impact show limitations, and we need new research directions and new perspectives to better capture changes in the carbon fluxes between the ecosystem and atmosphere [84]. Indeed, a long-standing debate around the C-neutrality of old-growth forests (and some of the sites become old-growth at the end of simulations) raises concerns, and debates will increase about the reduction of the sink capacity of aging forests, as assumed by Odum’s theory. As found in [7] (but see [85]), we also found that even a >200-year-old stand (as IT-Col in 2099) still has sink capacity. The annual NEE decrease (and much less the changes in the MSC), as shown by the model results, is certainly a function not only of climate but also of the inherent effects related to, e.g., biomass, both live and dead, accumulation (which led to increases in respiratory costs), and changes in, e.g., the forest structure (which led to decreases in the carbon assimilation), as stands age. Nevertheless, such effects on annual NEE are expected to be greatly exacerbated by climate change.

#### 4.3. Changes in NEE Dynamics under Different Climate Scenarios

Climate change impacts plant phenology by altering the start and end dates of the growing season, which influences when photosynthesis can begin and consequently affects C-fluxes [20,29,30,32,86].

A primary finding is that, regardless of the scenario analyzed, the number of days identified as a C-sink in evergreen forests increases significantly over time, indicating a consistent upward trend. Similarly, the number of days classified as a C-source decreases over the decades, showing a general reduction. The second finding is that for evergreen forests, the DoY to C-sink tends to increase (indicating a forward shift in the year when the system becomes a sink), and the DoY to C-source decreases (indicating a backward shift in the year when the system becomes a source), aligning with the overall trend of fewer C-sink days and more C-source days over time. In contrast, deciduous forests maintain a relatively stable number of C-sink (and C-source) days throughout the century, reflecting a steady DoY when the system becomes a C-sink (or a C-source), despite an anticipated beginning of the growing season but compensated by higher respiration rates. Indeed, for the deciduous, the 3D-CMCC-FEM simulates the bud-brake through a thermic sum function and leaf and fine root development (and the relative growth respiration peak) in a well-defined and short time during spring [36]. Conversely, for the evergreens, leaf and fine root growth development is spread all over the spring. At the same time, leaf fall in the deciduous starts under certain hours of solar radiation, and thus, this is not under the control of climate, while in the evergreen, it happens all over the year, and under the control of climate, balanced by incoming photosynthates for new leaves and fine roots. Ultimately, deciduous spring C-sink capacity is counterbalanced by high C-emissions mainly because of growth respiration in spring. Such behavior is different for evergreens, which lengthen their C-sink capacity during spring. However, the lengthening of the growing season does not automatically mirror an increase in the net sink capacity because  $R_{eco}$  shows an increase much more than GPP. It is generally acknowledged that the changing temperature response of respiration transforms forests from C-sinks to C-sources [26,76,87–89], while the stability of the carbon sink/source dynamics over the decades for deciduous ecosystems is a relatively recent finding. The DoY for deciduous forests remains unchanged because the earlier start of the growing season, triggered by rising temperatures, is balanced by an earlier increase in respiration. This compensates for the earlier rise in GPP at the level of NEE. Overall, the lack of change in the number of C-sink (and C-source) days across decades and the reduction in the NEE suggest that, over the long term, deciduous forests are more efficient in using photosynthates compared to evergreen forests [90,91].

#### 5. Limitations

The modeling framework presented has certain limitations that must be acknowledged. First, we deliberately decided to not simulate the effects of anthropogenic disturbances, e.g., forest management, nor the ones from natural disturbances caused by climate change, such as windstorms, forest fires, and insect outbreaks, so as to concentrate on the effects of climate change alone and avoid these potentially confounding effects. Climate extreme events are presumed to be incorporated into the climate scenarios used to drive the model and are, therefore, somewhat already accounted for. Moreover, indirect changes due to climate change in key factors like nitrogen deposition, phosphorus, or ozone—which could potentially amplify or mitigate our findings—were not evaluated. Nonetheless, some research (e.g., [92]) indicates that this issue might not be universally relevant. These studies highlight the strong response of various tree species to CO<sub>2</sub> fertilization across different levels of nutrient availability. Lastly, the potential for species migration to and from the study areas was not considered. However, such dynamics might require longer timescales than those covered in this study and it is unlikely (although still possible in theory) that species composition may completely change throughout the simulations.

## 6. Conclusions

The MSC metric is an interesting and descriptive metric that associates phenology and carbon partitioning allocation within forest stands. Climate change impacts both the phenology, by changing the date for the beginning and the end of the growing season, and the ecosystem carbon allocation.

We applied the process-based forest model 3D-CMCC-FEM to evaluate the potential modifications on the ability of different forest stands to absorb or emit carbon throughout the season and across the years up to 2100. Before that, we validated the model under current climate conditions and found a robust predictive ability of 3D-CMCC-FEM in estimating NEE, GPP, and  $R_{eco}$  across different timeframes, forest species, and climates.

The analysis of the mean NEE seasonal cycle under different climate change scenarios presents a consistent reduction in the forest stand capabilities to act as a carbon sink in the coming years, across all study sites, and climate change scenarios. The reduction in NEE ability has different intensities of climate impacts across these scenarios. The RCP 2.6 scenario demonstrates a less pronounced reduction in NEE compared to the RCP 6.0 scenario. This disparity primarily stems from variations in key variables, such as the differing rates of temperature increase between the two scenarios, as well as the  $CO_2$  fertilization effect, while, in all sites, age effects depend on the age at the beginning of the simulations. The reduction in C-sink capabilities is mainly notable in evergreen forests, which exhibit a higher decrease in NEE compared to deciduous forest sites.

Finally, we found that the number of days as a C-sink in evergreen forests increases over the years, indicating a consistent upward trend. Conversely, the number of days as a C-source decreases over the decades, showing a general reduction. This statement aligns with the forward shift of DoY to C-sink, and the backward shift of DoY to C-source. In contrast, deciduous forests maintain a relatively stable number of C-sink (and C-source) days throughout the century, reflecting a fixed DoY when the system becomes a C-sink (or a C-source). The DoY for deciduous forests remains constant as the earlier onset of the growing season, driven by warming temperatures, is offset by an earlier uptick in respiration. Decades pass with little change in the number of days as a C-sink (and a C-source), alongside a decrease in NEE. This indicates that deciduous forests, over the long haul, demonstrate greater efficiency in utilizing photosynthates when compared to evergreen forests.

**Author Contributions:** M.M.: Data curation, Formal analysis, Investigation, Writing—original draft, Writing—review and editing; E.V.: Data curation, Writing—original draft, Writing—review and editing; A.C.: Conceptualization, Formal analysis, Investigation, Writing—original draft, Writing—review and editing. All authors have read and agreed to the published version of the manuscript.

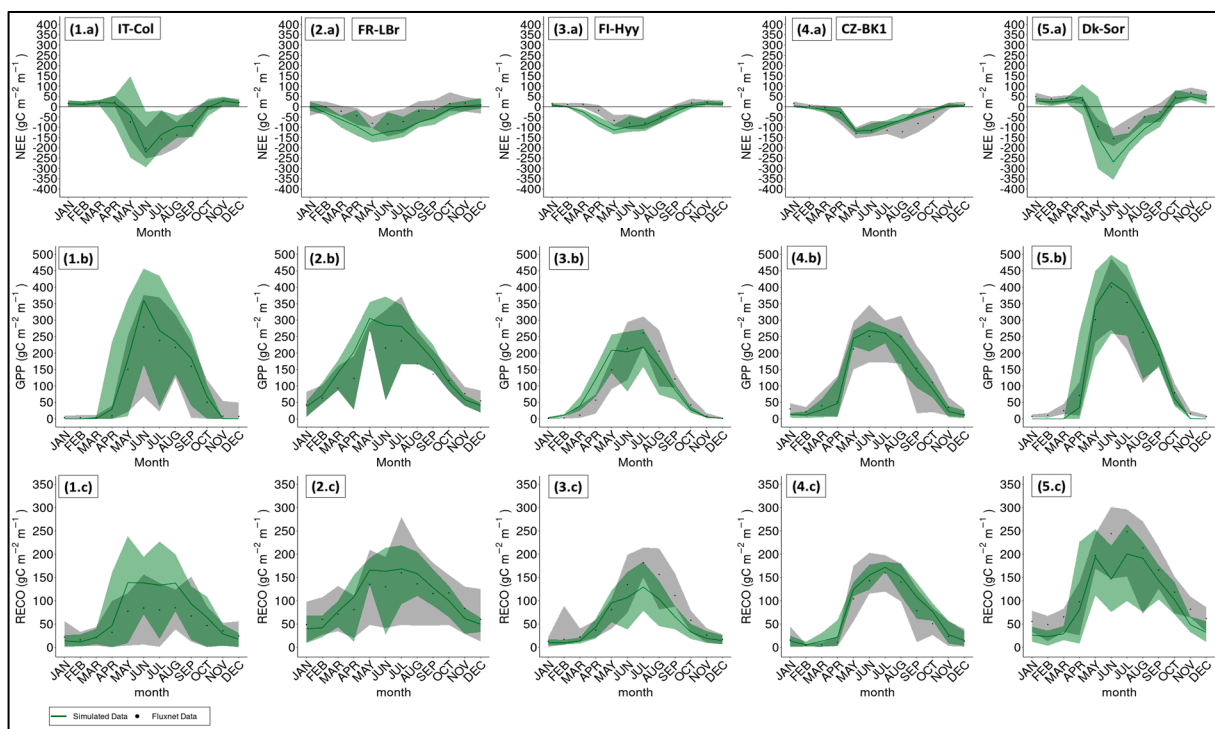
**Funding:** OptForEU Horizon Europe research and innovation programme under grant agreement No. 101060554; National Recovery and Resilience Plan (NRRP), Mission 4 Component 2 Investment 1.4—Call for tender No. 3138 of 16 December 2021, rectified by Decree n.3175 of 18 December 2021 of Italian Ministry of University and Research funded by the European Union—NextGenerationEU under award number: Project code CN\_00000033, Concession Decree No. 1034 of 17 June 2022 adopted by the Italian Ministry of University and Research, CUP B83C22002930006; Project title “National Biodiversity Future Centre—NBFC”; A.C. also acknowledges funding from the MIUR Project (PRIN 2020)—Research Projects of National Relevance funded by the Italian Ministry of University and Research entitled: “Unraveling interactions between WATER and carbon cycles during drought and their impact on water resources and forest and grassland ecosystems in the Mediterranean climate” (WATERSTEM, project number: 20202WF53Z), and “WAFER” at CNR (Consiglio Nazionale delle Ricerche); A.C. and E.V. also acknowledge funding from the MIUR Project (PRIN 2020)—Research Projects of National Relevance funded by the Italian Ministry of University and Research entitled: “Multi-scale observations to predict Forest response to pollution and climate change” (MULTIFOR, project number: 2020E52THS).

**Data Availability Statement:** The 3D-CMCC-FEM model code is publicly available and can be found on the GitHub platform at: <https://github.com/Forest-Modelling-Lab/3D-CMCC-FEM> (accessed on 24 June 2024). The 3D-CMCC-FEM output data used in this work can be downloaded at: <https://zenodo.org/records/11124413> (accessed on 7 May 2024). Correspondence and requests for additional materials should be addressed to the corresponding author.

**Acknowledgments:** We are thankful to D. Dalmonech for supporting data preparation and analysis. We also thank the ISIMIP project and the COST Action FP1304 PROFOUND (Towards Robust Projections of European Forests under Climate Change), supported by COST (European Cooperation in Science and Technology) for providing us the climate historical and future scenarios and site data used in this work. This work used eddy covariance data acquired and shared by the FLUXNET community, including these networks: AmeriFlux, AfriFlux, AsiaFlux, CarboAfrica, CarboEurope-IP, CarboItaly, CarboMont, ChinaFlux, Fluxnet-Canada, GreenGrass, ICOS, KoFlux, LBA, NECC, OzFlux-TERN, TCOS-Siberia, and USCCC. The ERA-Interim reanalysis data are provided by ECMWF and processed by LSCE. The FLUXNET eddy covariance data processing and harmonization was carried out by the European Fluxes Database Cluster, AmeriFlux Management Project, and Fluxdata Project of FLUXNET, with the support of CDIAC and ICOS Ecosystem Thematic Center, and the OzFlux, ChinaFlux, and AsiaFlux offices. We acknowledge the World Climate Research Programme's Working Group on Coupled Modelling, which is responsible for CMIP, and we thank the respective climate modeling groups for producing and making available their model output. The U.S. Department of Energy's Program for Climate Model Diagnosis and Intercomparison at Lawrence Livermore National Laboratory provides coordinating support for CMIP, and led the development of software infrastructure in partnership with the Global Organization for Earth System Science Portals.

**Conflicts of Interest:** The authors declare no conflicts of interest.

## Appendix A



**Figure A1.** The green lines represent the modeled (a) NEE, (b) GPP, and (c)  $R_{eco}$  amounts ( $gC\ m^{-2}\ month^{-1}$ ) per month for the five selected case studies (i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-Lbr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor) compared to relative observed data (depicted as black dots) from the Fluxnet2015 Dataset (Pastorello et al., 2020 [56]). The lower and upper lines of the shaded area represent, respectively, the minimum and maximum values of the observed and modeled datasets considered.



**Table A1.** Summary of the statistics between simulated and measured NEE from the Fluxnet2015 Dataset (Pastorello et al., 2020 [56]), calculated on the 5 cases studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). The table shows the daily and monthly values for Person’s Coefficient ( $r$ —dimensionless), Relative Mean Bias (RMB—dimensionless), Normalized Root Mean Square Error (NRMSE—dimensionless), Modeling Efficiency (ME—dimensionless), and Mean Absolute Error (MAE,  $\text{gC m}^{-2} \text{ time}^{-1}$ ).

	IT-Col	FR-LBr	FI-Hyy	CZ-Bk1	DK-Sor
Daily NEE					
$r$	0.96	0.93	0.93	0.85	0.97
RMB	0.07	−0.86	−0.48	0.14	−0.46
NRMSE	−0.57	−1.56	−1.54	1.39	−6.59
ME	0.92	−0.10	0.58	0.69	0.46
MAE	0.67	1.20	0.67	1.15	1.42
Monthly NEE					
$r$	0.98	0.95	0.93	0.92	0.99
RMB	0.07	−0.88	−0.49	0.21	−0.46
NRMSE	0.99	−1.55	−1.52	−0.50	−6.14
ME	0.95	−0.15	0.55	0.79	0.55
MAE	11.42	33.39	19.63	19.12	37.78

**Table A2.** Summary of the statistics between simulated and measured GPP from the Fluxnet2015 Dataset (Pastorello et al., 2020 [56]), calculated on the 5 cases studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-Lbr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). The table shows the daily and monthly values for Person’s Coefficient ( $r$ —dimensionless), Relative Mean Bias (RMB—dimensionless), Normalized Root Mean Square Error (NRMSE—dimensionless), Modeling Efficiency (ME—dimensionless), and Mean Absolute Error (MAE,  $\text{gC m}^{-2} \text{ time}^{-1}$ ).

	IT-Col	FR-LBr	FI-Hyy	CZ-Bk1	DK-Sor
Daily GPP					
$r$	0.99	0.95	0.93	0.95	0.99
RMB	0.15	0.55	0.01	−0.14	0.00
NRMSE	0.34	0.42	0.39	1.21	0.19
ME	0.91	0.28	0.86	0.85	0.97
MAE	0.91	1.51	0.91	0.97	0.75
Monthly GPP					
$r$	0.99	0.96	0.93	0.99	1.00
RMB	0.15	0.52	0.01	−0.06	0.00
NRMSE	0.31	0.40	0.38	0.15	0.16
ME	0.92	0.34	0.86	0.97	0.97
MAE	20.68	41.69	26.05	14.25	20.26

**Table A3.** Summary of the statistics between simulated and measured  $R_{\text{eco}}$  from the Fluxnet2015 Dataset (Pastorello et al., 2020 [56]), calculated on the 5 cases studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). The table shows the daily and monthly values for Person’s Coefficient ( $r$ —dimensionless), Relative Mean Bias (RMB—dimensionless), Normalized Root Mean Square Error (NRMSE—dimensionless), Modeling Efficiency (ME—dimensionless), and Mean Absolute Error (MAE,  $\text{gC m}^{-2} \text{ time}^{-1}$ ).

	IT-Col	FR-LBr	FI-Hyy	CZ-Bk1	DK-Sor
Daily $R_{\text{eco}}$					
$r$	0.90	0.89	0.94	0.86	0.89
RMB	0.82	0.16	−0.30	0.30	−0.42
NRMSE	0.72	0.24	0.41	1.16	0.35
ME	−1.17	0.50	0.75	0.63	0.62
MAE	1.14	0.71	0.74	0.88	1.29

**Table A3.** *Cont.*

	IT-Col	FR-LBr	FI-Hyy	CZ-Bk1	DK-Sor
Monthly $R_{eco}$					
$r$	0.99	0.94	0.95	0.99	0.95
RMB	0.77	0.14	−0.29	0.19	−0.42
NRMSE	0.83	0.20	0.39	0.24	0.29
ME	−0.64	0.68	0.77	0.93	0.71
MAE	25.42	18.01	21.02	11.61	31.82

**Table A4.** Summary of the NEE variation (%) from the first decade of simulation (1999–2009), considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones. Note that negative values indicate that NEE becomes less negative, e.g., −100% indicates a reduction in the negative values of NEE.

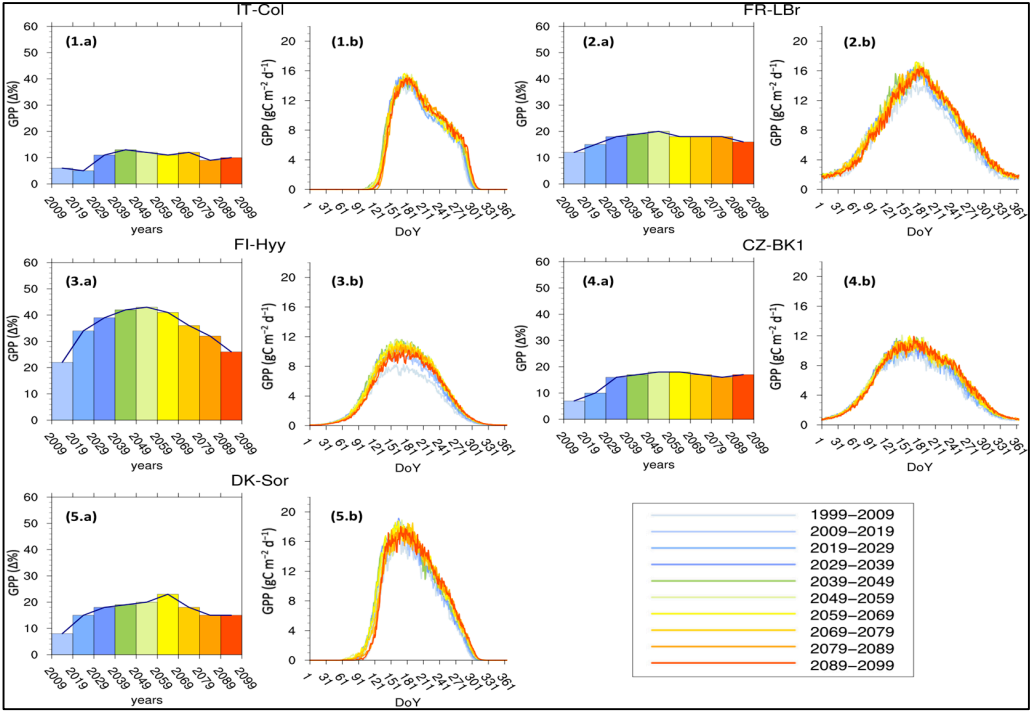
NEE					
year	RCP 2.6				
	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
2009–2019	− <b>37%</b>	−19%	− <b>4%</b>	−24%	−9%
2019–2029	− <b>64%</b>	−35%	− <u>16%</u>	−44%	−14%
2029–2039	− <b>70%</b>	−31%	−27%	−53%	− <u>22%</u>
2039–2049	− <b>76%</b>	−39%	−39%	−59%	− <u>29%</u>
2049–2059	− <b>71%</b>	−48%	−47%	−58%	− <u>33%</u>
2059–2069	− <b>79%</b>	−59%	−58%	−65%	− <u>32%</u>
2069–2079	− <b>75%</b>	−61%	−66%	−70%	− <u>37%</u>
2079–2089	−76%	−69%	− <b>76%</b>	−75%	− <u>44%</u>
2089–2099	−74%	−68%	− <b>95%</b>	−72%	− <u>43%</u>
year	RCP 6.0				
	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
2009–2019	− <b>35%</b>	−11%	0%	−25%	−11%
2019–2029	− <b>60%</b>	−27%	−11%	−43%	−14%
2029–2039	− <b>74%</b>	−36%	− <u>23%</u>	−52%	−26%
2039–2049	− <b>76%</b>	−43%	−36%	−57%	− <u>34%</u>
2049–2059	− <b>75%</b>	−44%	−46%	−61%	− <u>31%</u>
2059–2069	− <b>75%</b>	−43%	−57%	−64%	− <u>23%</u>
2069–2079	− <b>72%</b>	−54%	−69%	−69%	− <u>30%</u>
2079–2089	−76%	−59%	− <b>79%</b>	−75%	− <u>34%</u>
2089–2099	−71%	−60%	−87%	−68%	− <u>36%</u>

**Table A5.** Summary of the GPP variation (%) from the first decade of simulation (1999–2009), considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones.

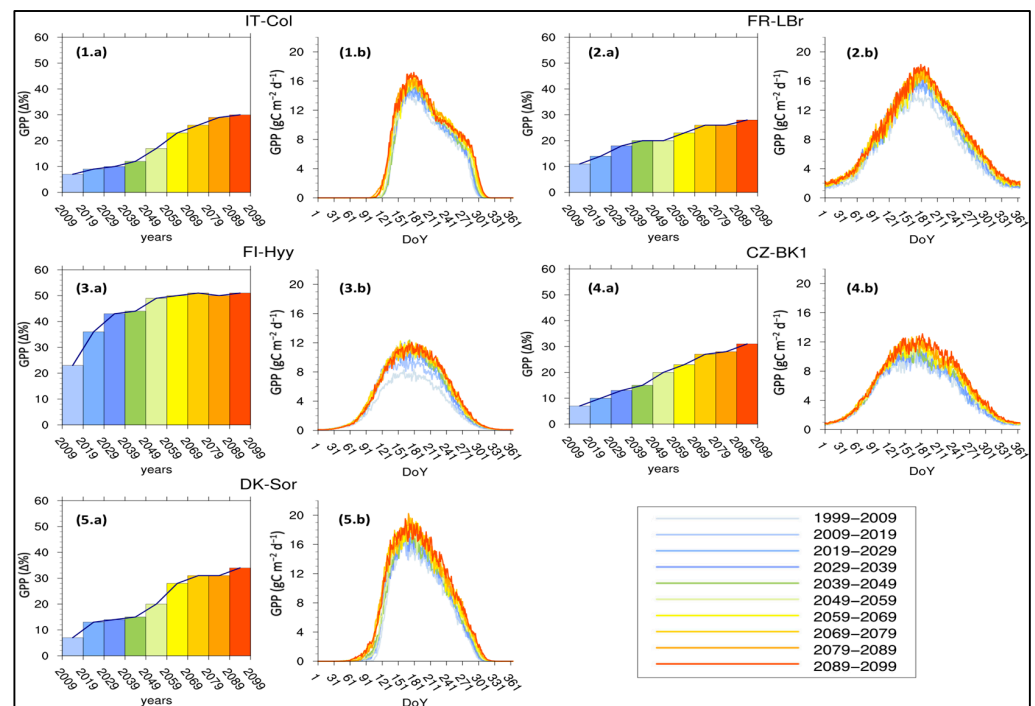
GPP					
year	RCP 2.6				
	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
2009–2019	7%	6%	<b>22%</b>	12%	8%
2019–2029	10%	<u>5%</u>	<b>34%</b>	15%	15%
2029–2039	16%	<u>11%</u>	<b>39%</b>	18%	18%
2039–2049	17%	<u>13%</u>	<b>42%</b>	19%	19%
2049–2059	18%	<u>12%</u>	<b>43%</b>	20%	20%
2059–2069	18%	<u>11%</u>	<b>41%</b>	18%	23%
2069–2079	17%	<u>12%</u>	<b>36%</b>	18%	18%
2079–2089	16%	<u>9%</u>	<b>32%</b>	18%	15%
2089–2099	17%	<u>10%</u>	<b>26%</b>	16%	15%

Table A5. Cont.

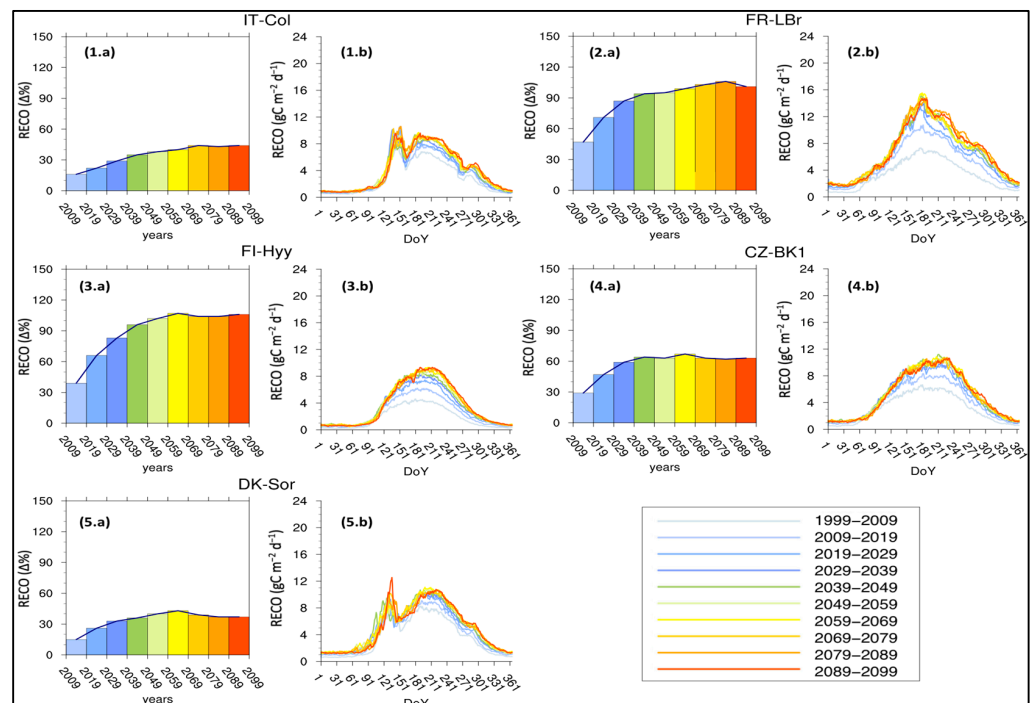
GPP					
year	RCP 6.0				
	CZ-BK1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
2009–2019	7%	7%	23%	11%	7%
2019–2029	10%	9%	36%	14%	13%
2029–2039	13%	10%	43%	18%	14%
2039–2049	15%	12%	44%	20%	15%
2049–2059	20%	17%	49%	20%	20%
2059–2069	23%	23%	50%	23%	28%
2069–2079	27%	26%	51%	26%	31%
2079–2089	28%	29%	50%	26%	31%
2089–2099	31%	30%	51%	28%	34%



**Figure A2.** Ten-year average GPP seasonal cycle under the RCP 2.6 climate scenario for 5 case studies selected, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. The histograms (a) represent the annual GPP variation (%) from the first decade taken as a benchmark of simulation (1999–2009). The xy plots (b) show the Mean Seasonal GPP Cycle of monthly values ( $\text{gC m}^{-2} \text{day}^{-1}$ ).

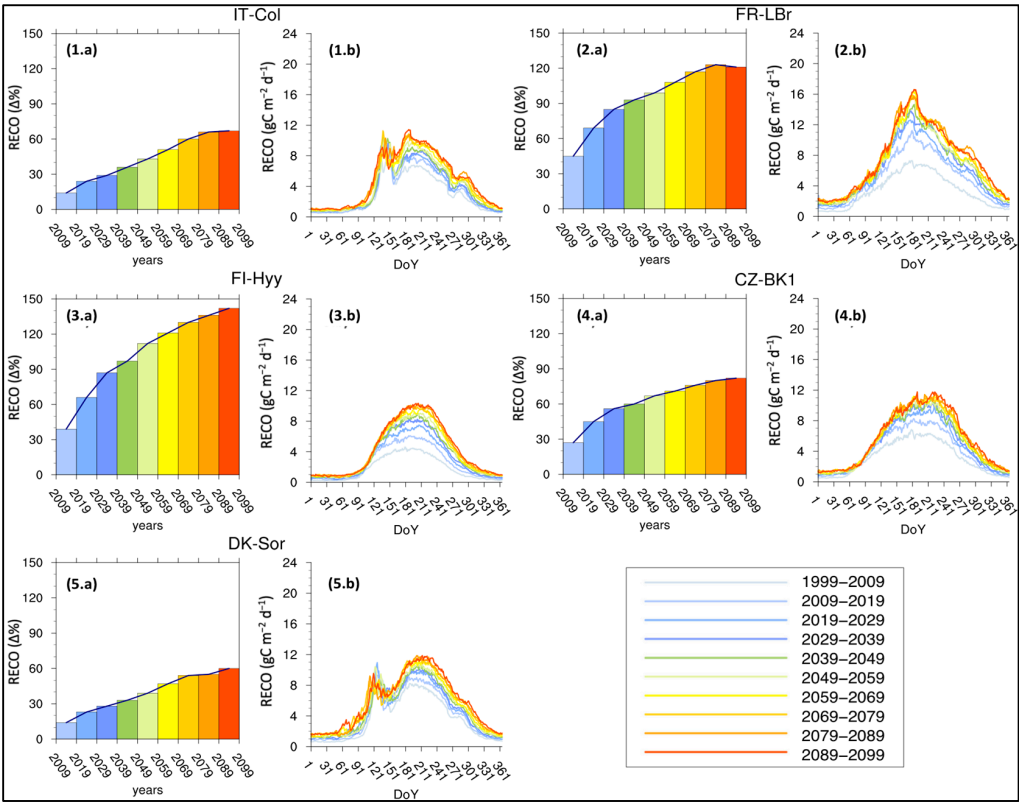


**Figure A3.** Ten-year average GPP seasonal cycle under the RCP 6.0 climate scenario for 5 case studies selected, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. The histograms (a) represent the annual GPP variation (%) from the first decade taken as a benchmark of simulation (1999–2009). The xy plots (b) show the Mean Seasonal GPP Cycle of monthly values ( $\text{gC m}^{-2} \text{ day}^{-1}$ ).



**Figure A4.** Ten-year average  $\text{Reco}$  seasonal cycle under the RCP 2.6 climate scenario for 5 case studies selected, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. The histograms (a) represent the annual  $\text{Reco}$  variation (%) from the first decade taken as a benchmark of simulation (1999–2009). The xy plots (b) show the Mean Seasonal  $\text{Reco}$  Cycle of monthly values ( $\text{gC m}^{-2} \text{ day}^{-1}$ ).





**Figure A5.** Ten-year average  $R_{eco}$  seasonal cycle under the RCP 6.0 climate scenario for 5 case studies selected, i.e., (1) Collelongo—IT-Col, (2) Le Bray—FR-LBr, (3) Hyytiälä—FI-Hyy, (4) Bílý Kříž—CZ-Bk1, and (5) Sorø—DK-Sor. The histograms (a) represent the annual  $R_{eco}$  variation (%) from the first decade taken as a benchmark of simulation (1999–2009). The xy plots (b) show the Mean Seasonal  $R_{eco}$  Cycle of monthly values ( $gC\ m^{-2}\ day^{-1}$ ).

**Table A6.** Summary of the  $R_{eco}$  changes (%) from the first decade of simulation (1999–2009), considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones.

$R_{eco}$					
RCP 2.6					
year	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
2009–2019	29%	16%	39%	<b>47%</b>	15%
2019–2029	47%	<u>22%</u>	66%	<b>71%</b>	26%
2029–2039	59%	29%	83%	<b>87%</b>	33%
2039–2049	64%	35%	<b>96%</b>	94%	36%
2049–2059	63%	<u>38%</u>	<b>102%</b>	95%	40%
2059–2069	67%	40%	<b>107%</b>	99%	43%
2069–2079	63%	44%	<b>104%</b>	103%	39%
2079–2089	62%	43%	104%	<b>106%</b>	<u>37%</u>
2089–2099	63%	44%	<b>106%</b>	101%	<u>37%</u>
RCP 6.0					
year	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
2009–2019	27%	14%	39%	<b>45%</b>	14%
2019–2029	45%	24%	66%	<b>69%</b>	<u>23%</u>
2029–2039	56%	29%	<b>87%</b>	85%	<u>28%</u>
2039–2049	60%	36%	<b>97%</b>	93%	33%
2049–2059	67%	43%	<b>112%</b>	99%	39%
2059–2069	71%	51%	<b>121%</b>	108%	47%
2069–2079	76%	60%	<b>130%</b>	117%	<u>54%</u>
2079–2089	80%	66%	<b>136%</b>	123%	55%
2089–2099	82%	67%	<b>142%</b>	121%	<u>60%</u>

**Table A7.** Summary of the number of the days in a year (N. days year<sup>−1</sup>) as C-sink, considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones within each forest stand.

Days as C-Sink					
year	CZ-Bk1	RCP 2.6			
		IT-Col	FI-Hyy	FR-Lbr	DK-Sor
1999–2009	<u>44</u>	221	<u>107</u>	0	212
2009–2019	126	223	135	<u>25</u>	208
2019–2029	176	223	156	82	<u>206</u>
2029–2039	182	217	170	85	216
2039–2049	195	<u>216</u>	171	103	217
2049–2059	189	220	182	102	218
2059–2069	<b>199</b>	220	199	108	216
2069–2079	189	217	201	109	219
2079–2089	194	<b>225</b>	208	<b>129</b>	222
2089–2099	193	219	<b>218</b>	117	<b>228</b>

year	CZ-Bk1	RCP 6.0			
		IT-Col	FI-Hyy	FR-Lbr	DK-Sor
1999–2009	<u>47</u>	221	<u>107</u>	0	215
2009–2019	108	223	134	<u>26</u>	216
2019–2029	172	222	155	71	218
2029–2039	195	223	164	87	216
2039–2049	194	<b>224</b>	180	96	220
2049–2059	190	220	181	95	<b>222</b>
2059–2069	<b>199</b>	212	201	101	216
2069–2079	<b>199</b>	211	197	111	<u>214</u>
2079–2089	193	213	206	<b>114</b>	<u>214</u>
2089–2099	188	<u>206</u>	<b>217</b>	112	<b>222</b>

**Table A8.** Summary of number of the days in a year (n. days year<sup>−1</sup>) source, considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones within each forest stand.

Days as C-Source					
year	CZ-Bk1	RCP 2.6			
		IT-Col	FI-Hyy	FR-LBr	DK-Sor
1999–2009	<b>321</b>	144	<b>258</b>	<b>365</b>	153
2009–2019	239	142	230	340	157
2019–2029	189	142	209	283	<b>159</b>
2029–2039	183	148	195	280	149
2039–2049	<u>170</u>	<b>149</b>	194	262	148
2049–2059	176	145	183	263	147
2059–2069	166	145	166	257	149
2069–2079	176	148	164	256	146
2079–2089	171	<u>140</u>	157	<u>236</u>	143
2089–2099	172	146	<u>147</u>	248	<u>137</u>

Table A8. Cont.

Days as C-Source					
year	CZ-Bk1	RCP 6.0			
		IT-Col	FI-Hyy	FR-LBr	DK-Sor
1999–2009	<b>318</b>	144	<b>258</b>	<b>365</b>	150
2009–2019	257	142	231	339	149
2019–2029	193	143	210	294	147
2029–2039	170	142	201	278	149
2039–2049	171	<u>141</u>	185	269	145
2049–2059	175	145	184	270	<u>143</u>
2059–2069	<u>166</u>	153	164	264	149
2069–2079	<u>166</u>	154	168	254	<b>151</b>
2079–2089	172	152	159	<u>251</u>	<b>151</b>
2089–2099	177	<b>159</b>	<u>148</u>	253	<u>143</u>

**Table A9.** Summary of the changes in the source/sink DoY (Day of Year), considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones within each forest stand. Data are missing for some intervals because of filtering and data removal to avoid pulsing artifacts, e.g., the ‘Birch effect’ (see Section 2).

DoY to C-Sink					
year	CZ-Bk1	RCP 2.6			
		IT-Col	FI-Hyy	FR-LBr	DK-Sor
1999–2009	<u>2</u>	141	<u>30</u>	-	125
2009–2019	13	<b>150</b>	35	<u>1</u>	118
2019–2029	<b>31</b>	142	<b>50</b>	2	116
2029–2039	18	137	-	-	119
2039–2049	24	<u>136</u>	45	11	117
2049–2059	22	139	47	8	120
2059–2069	26	139	<b>50</b>	7	<u>115</u>
2069–2079	25	<u>136</u>	44	9	118
2079–2089	21	144	48	<b>16</b>	123
2089–2099	21	138	<b>50</b>	8	<b>127</b>

year	CZ-Bk1	RCP 6.0			
		IT-Col	FI-Hyy	FR-LBr	DK-Sor
1999–2009	-	142	<u>30</u>	-	<b>129</b>
2009–2019	<u>11</u>	<b>146</b>	38	<u>1</u>	124
2019–2029	13	141	40	-	123
2029–2039	25	144	40	<u>1</u>	120
2039–2049	14	144	53	-	120
2049–2059	21	140	46	<u>1</u>	120
2059–2069	<b>32</b>	132	53	<b>8</b>	111
2069–2079	28	133	52	<b>8</b>	<u>107</u>
2079–2089	27	130	51	4	111
2089–2099	31	<u>127</u>	<b>59</b>	<u>1</u>	112

**Table A10.** Summary of the changes in the sink/source DoY (Day of Year), considering both climate change scenarios (RCP 2.6 and 6.0) for 5 case studies selected (i.e., Collelongo—IT-Col, Le Bray—FR-LBr, Hyytiälä—FI-Hyy, Bílý Kříž—CZ-Bk1, and Sorø—DK-Sor). In bold values where changes were the highest between the decades while underlined the lowest ones within each forest stand. Data are missing for some intervals because of filtering and data removal to avoid pulsing artifacts, e.g., the ‘Birch effect’ (see Section 2).

DoY to C-Source					
year	RCP 2.6				
	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
1999–2009	<b>329</b>	285	<b>285</b>	-	<b>280</b>
2009–2019	280	<b>286</b>	273	-	277
2019–2029	203	<u>284</u>	265	<b>277</b>	275
2029–2039	198	285	243	273	271
2039–2049	200	285	258	274	<u>267</u>
2049–2059	<u>194</u>	<u>284</u>	244	268	268
2059–2069	189	<u>284</u>	221	265	269
2069–2079	201	<u>284</u>	214	267	<u>267</u>
2079–2089	196	<u>284</u>	207	<u>245</u>	272
2089–2099	200	<u>284</u>	<u>197</u>	255	268
year	RCP 6.0				
	CZ-Bk1	IT-Col	FI-Hyy	FR-LBr	DK-Sor
1999–2009	<b>312</b>	285	<b>288</b>	-	<b>277</b>
2009–2019	264	285	275	-	273
2019–2029	205	284	252	280	270
2029–2039	197	284	245	<b>283</b>	269
2039–2049	183	283	245	273	269
2049–2059	202	283	235	273	268
2059–2069	195	285	226	271	269
2069–2079	203	<b>287</b>	227	259	<u>262</u>
2079–2089	193	<u>282</u>	<u>208</u>	<u>258</u>	263
2089–2099	204	286	212	260	266

## References

1. Ciais, P.; Sabine, C.; Bala, G.; Peters, W. Carbon and other biogeochemical cycles. In *Climate Change 2013 the Physical Science Basis: Working Group I Contribution to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*; Cambridge University Press: Cambridge, UK, 2013; Volume 9781107057, pp. 465–570. ISBN 9781107415324.
2. Anav, A.; Friedlingstein, P.; Beer, C.; Ciais, P.; Harper, A.; Jones, C.; Murray-Tortarolo, G.; Papale, D.; Parazoo, N.C.; Peylin, P.; et al. Spatiotemporal patterns of terrestrial gross primary production: A review. *Rev. Geophys.* **2015**, *53*, 785–818. [\[CrossRef\]](#)
3. Liang, M.C.; Laskar, A.H.; Barkan, E.; Newman, S.; Thiemens, M.H.; Rangarajan, R. New constraints of terrestrial and oceanic global gross primary productions from the triple oxygen isotopic composition of atmospheric CO<sub>2</sub> and O<sub>2</sub>. *Sci. Rep.* **2023**, *13*, 1–10. [\[CrossRef\]](#) [\[PubMed\]](#)
4. Intergovernmental Panel on Climate Change. *Climate Change 2021—The Physical Science Basis*; Cambridge University Press: Cambridge, UK, 2023. [\[CrossRef\]](#)
5. Quéré, C.; Andrew, R.; Friedlingstein, P.; Sitch, S.; Hauck, J.; Pongratz, J.; Pickers, P.; Ivar Korsbakken, J.; Peters, G.; Canadell, J.; et al. Global Carbon Budget 2018. *Earth Syst. Sci. Data* **2018**, *10*, 2141–2194. [\[CrossRef\]](#)
6. Rödenbeck, C.; Zaehle, S.; Keeling, R.; Heimann, M. How does the terrestrial carbon exchange respond to inter-Annual climatic variations? A quantification based on atmospheric CO<sub>2</sub> data. *Biogeosciences* **2018**, *15*, 2481–2498. [\[CrossRef\]](#)
7. Luyssaert, S.; Schulze, E.D.; Börner, A.; Knohl, A.; Hessenmöller, D.; Law, B.E.; Ciais, P.; Grace, J. Old-growth forests as global carbon sinks. *Nature* **2008**, *455*, 213–215. [\[CrossRef\]](#) [\[PubMed\]](#)
8. Collalti, A.; Ibrom, A.; Stockmarr, A.; Cescatti, A.; Alkama, R.; Fernández-Martínez, M.; Matteucci, G.; Sitch, S.; Friedlingstein, P.; Ciais, P.; et al. Forest production efficiency increases with growth temperature. *Nat. Commun.* **2020**, *11*, 5322. [\[CrossRef\]](#) [\[PubMed\]](#)
9. Amiro, B.D.; Barr, A.G.; Barr, J.G.; Black, T.A.; Bracho, R.; Brown, M.; Chen, J.; Clark, K.L.; Davis, K.J.; Desai, A.R.; et al. Ecosystem carbon dioxide fluxes after disturbance in forests of North America. *J. Geophys. Res. Biogeosci.* **2010**, *115*, 237–251. [\[CrossRef\]](#)
10. Collalti, A.; Trotta, C.; Keenan, T.F.; Ibrom, A.; Bond-Lamberty, B.; Grote, R.; Vicca, S.; Reyer, C.P.O.; Migliavacca, M.; Veroustraete, F.; et al. Thinning Can Reduce Losses in Carbon Use Efficiency and Carbon Stocks in Managed Forests Under Warmer Climate. *J. Adv. Model. Earth Syst.* **2018**, *10*, 2427–2452. [\[CrossRef\]](#)

11. Dalmonech, D.; Marano, G.; Amthor, J.S.; Cescatti, A.; Lindner, M.; Trotta, C.; Collalti, A. Feasibility of enhancing carbon sequestration and stock capacity in temperate and boreal European forests via changes to management regimes. *Agric. For. Meteorol.* **2022**, *327*, 109203. [\[CrossRef\]](#)
12. Tramontana, G.; Migliavacca, M.; Jung, M.; Reichstein, M.; Keenan, T.F.; Camps-Valls, G.; Ogee, J.; Verrelst, J.; Papale, D. Partitioning net carbon dioxide fluxes into photosynthesis and respiration using neural networks. *Glob. Chang. Biol.* **2020**, *26*, 5235–5253. [\[CrossRef\]](#)
13. Santini, M.; Collalti, A.; Valentini, R. Climate change impacts on vegetation and water cycle in the Euro-Mediterranean region, studied by a likelihood approach. *Reg. Environ. Change* **2014**, *14*, 1405–1418. [\[CrossRef\]](#)
14. Noce, S.; Collalti, A.; Valentini, R.; Santini, M. Hot spot maps of forest presence in the Mediterranean basin. *iForest* **2016**, *9*, 766–774. [\[CrossRef\]](#)
15. Lionello, P.; Scarascia, L. The relation between climate change in the Mediterranean region and global warming. *Reg. Environ. Change* **2018**, *18*, 1481–1493. [\[CrossRef\]](#)
16. FAO. State of Mediterranean Forests (SoFMF), Concept Paper; Marseille, 2011. Available online: <http://www.fao.org/docrep/013/ma723e/ma723e00.pdf> (accessed on 10 May 2024).
17. Collalti, A.; Thornton, P.E.; Cescatti, A.; Rita, A.; Borghetti, M.; Nolè, A.; Trotta, C.; Ciais, P.; Matteucci, G. The sensitivity of the forest carbon budget shifts across processes along with stand development and climate change. *Ecol. Appl.* **2019**, *29*, e01837. [\[CrossRef\]](#) [\[PubMed\]](#)
18. Friedlingstein, P.; Bopp, L.; Ciais, P.; Dufresne, J.L.; Fairhead, L.; LeTreut, H.; Monfray, P.; Orr, J. Positive feedback between future climate change and the carbon cycle. *Geophys. Res. Lett.* **2001**, *28*, 1543–1546. [\[CrossRef\]](#)
19. Chen, C.; Riley, W.J.; Prentice, I.C.; Keenan, T.F. CO<sub>2</sub> fertilization of terrestrial photosynthesis inferred from site to global scales. *Proc. Natl. Acad. Sci. USA* **2022**, *119*, e2115627119. [\[CrossRef\]](#)
20. Keenan, T.F.; Prentice, I.C.; Canadell, J.G.; Williams, C.A.; Wang, H.; Raupach, M.; Collatz, G.J. Recent pause in the growth rate of atmospheric CO<sub>2</sub> due to enhanced terrestrial carbon uptake. *Nat. Commun.* **2016**, *7*, 13428. [\[CrossRef\]](#)
21. Wang, S.; Zhang, Y.; Ju, W.; Chen, J.M.; Ciais, P.; Cescatti, A.; Sardans, J.; Janssens, I.A.; Wu, M.; Berry, J.A.; et al. Recent global decline of CO<sub>2</sub> fertilization effects on vegetation photosynthesis. *Science* **2020**, *370*, 1295–1300. [\[CrossRef\]](#)
22. Yuan, Z.; Ali, A.; Jucker, T.; Ruiz-Benito, P.; Wang, S.; Jiang, L.; Wang, X.; Lin, F.; Ye, J.; Hao, Z.; et al. Multiple abiotic and biotic pathways shape biomass demographic processes in temperate forests. *Ecology* **2019**, *100*, e02650. [\[CrossRef\]](#)
23. Grossiord, C.; Buckley, T.N.; Cernusak, L.A.; Novick, K.A.; Poulter, B.; Siegwolf, R.T.W.; Sperry, J.S.; McDowell, N.G. Plant responses to rising vapor pressure deficit. *New Phytol.* **2020**, *226*, 1550–1566. [\[CrossRef\]](#)
24. D'Andrea, E.; Rezaie, N.; Prislan, P.; Gričar, J.; Collalti, A.; Muhr, J.; Matteucci, G. Frost and drought: Effects of extreme weather events on stem carbon dynamics in a Mediterranean beech forest. *Plant Cell Environ.* **2020**, *43*, 2365–2379. [\[CrossRef\]](#) [\[PubMed\]](#)
25. McDowell, N.G.; Allen, C.D.; Anderson-Teixeira, K.; Aukema, B.H.; Bond-Lamberty, B.; Chini, L.; Clark, J.S.; Dietze, M.; Grossiord, C.; Hanbury-Brown, A.; et al. Pervasive shifts in forest dynamics in a changing world. *Science* **2020**, *368*, eaaz9463. [\[CrossRef\]](#) [\[PubMed\]](#)
26. Hadden, D.; Grelle, A. Changing temperature response of respiration turns boreal forest from carbon sink into carbon source. *Agric. For. Meteorol.* **2016**, *223*, 30–38. [\[CrossRef\]](#)
27. Mitchard, E.T.A. The tropical forest carbon cycle and climate change. *Nature* **2018**, *559*, 527–534. [\[CrossRef\]](#) [\[PubMed\]](#)
28. Bastida, F.; Torres, I.F.; Andrés-Abellán, M.; Baldrian, P.; López-Mondéjar, R.; Větrovský, T.; Richnow, H.H.; Starke, R.; Ondoño, S.; García, C.; et al. Differential sensitivity of total and active soil microbial communities to drought and forest management. *Glob. Chang. Biol.* **2017**, *23*, 4185–4203. [\[CrossRef\]](#) [\[PubMed\]](#)
29. Gunderson, C.A.; Edwards, N.T.; Walker, A.V.; O'Hara, K.H.; Campion, C.M.; Hanson, P.J. Forest phenology and a warmer climate—Growing season extension in relation to climatic provenance. *Glob. Chang. Biol.* **2012**, *18*, 2008–2025. [\[CrossRef\]](#)
30. Richardson, A.D.; Keenan, T.F.; Migliavacca, M.; Ryu, Y.; Sonnentag, O.; Toomey, M. Climate change, phenology, and phenological control of vegetation feedbacks to the climate system. *Agric. For. Meteorol.* **2013**, *169*, 156–173. [\[CrossRef\]](#)
31. Keenan, T.F.; Gray, J.; Friedl, M.A.; Toomey, M.; Bohrer, G.; Hollinger, D.Y.; Munger, J.W.; O'Keefe, J.; Schmid, H.P.; Wing, I.S.; et al. Net carbon uptake has increased through warming-induced changes in temperate forest phenology. *Nat. Clim. Change* **2014**, *4*, 598–604. [\[CrossRef\]](#)
32. Peano, D.; Matera, S.; Collalti, A.; Alessandri, A.; Anav, A.; Bombelli, A.; Gualdi, S. Global Variability of Simulated and Observed Vegetation Growing Season. *J. Geophys. Res. Biogeosci.* **2019**, *124*, 3569–3587. [\[CrossRef\]](#)
33. De Kauwe, M.G.; Medlyn, B.E.; Zaehle, S.; Walker, A.P.; Dietze, M.C.; Wang, Y.P.; Luo, Y.; Jain, A.K.; El-Masri, B.; Hickler, T.; et al. Where does the carbon go? A model–data intercomparison of vegetation carbon allocation and turnover processes at two temperate forest free-air CO<sub>2</sub> enrichment sites. *New Phytol.* **2014**, *203*, 883–899. [\[CrossRef\]](#)
34. Vacchiano, G.; Magnani, F.; Collalti, A. Modeling Italian forests: State of the art and future challenges. *Iforest-Biogeosci. For.* **2012**, *5*, 113. [\[CrossRef\]](#)
35. Maréchaux, I.; Langerwisch, F.; Huth, A.; Bugmann, H.; Morin, X.; Reyer, C.P.O.; Seidl, R.; Collalti, A.; Dantas de Paula, M.; Fischer, R.; et al. Tackling unresolved questions in forest ecology: The past and future role of simulation models. *Ecol. Evol.* **2021**, *11*, 3746–3770. [\[CrossRef\]](#) [\[PubMed\]](#)



36. Collalti, A.; Marconi, S.; Ibrom, A.; Trotta, C.; Anav, A.; D'andrea, E.; Matteucci, G.; Montagnani, L.; Gielen, B.; Mammarella, I.; et al. Validation of 3D-CMCC Forest Ecosystem Model (v.5.1) against eddy covariance data for 10 European forest sites. *Geosci. Model Dev.* **2016**, *9*, 479–504. [\[CrossRef\]](#)
37. Collalti, A.; Biondo, C.; Buttafuoco, G.; Maesano, M.; Caloiero, T.; Lucà, F.; Pellicone, G.; Ricca, N.; Salvati, R.; Veltri, A.; et al. Simulation, calibration and validation protocols for the model 3D-CMCC-CNR-FEM: A case study in the Bonis' watershed (Calabria, Italy). *For.-Riv. Selvic. Ecol. For.* **2017**, *14*, 247–256. [\[CrossRef\]](#)
38. Collalti, A.; Tjoelker, M.G.; Hoch, G.; Mäkelä, A.; Guidolotti, G.; Heskell, M.; Petit, G.; Ryan, M.G.; Battipaglia, G.; Matteucci, G.; et al. Plant respiration: Controlled by photosynthesis or biomass? *Glob. Chang. Biol.* **2020**, *26*, 1739–1753. [\[CrossRef\]](#)
39. Collalti, A.; Dalmonech, D.; Vangi, E.; Marano, G.; Puchi, P.F.; Morichetti, M.; Saponaro, V.; Orrico, M.R.; Grieco, E. *Monitoring and Predicting Forest Growth and Dynamics*; CNR Edizioni: Rome, Italy, 2024; ISBN 978-88-8080-655-4. [\[CrossRef\]](#)
40. Marconi, S.; Chiti, T.; Nolè, A.; Valentini, R.; Collalti, A. The role of respiration in estimation of net carbon cycle: Coupling soil carbon dynamics and canopy turnover in a novel version of 3D-CMCC forest ecosystem model. *Forests* **2017**, *8*, 220. [\[CrossRef\]](#)
41. Dalmonech, D.; Vangi, E.; Chiesi, M.; Chirici, G.; Fibbi, L.; Giannetti, F.; Marano, G.; Massari, C.; Nolè, A.; Xiao, J.; et al. Regional estimates of gross primary production applying the Process-Based Model 3D-CMCC-FEM vs. Remote-Sensing multiple datasets. *Eur. J. Remote Sens.* **2024**, *57*, 2301657. [\[CrossRef\]](#)
42. Mahnken, M.; Cailleret, M.; Collalti, A.; Trotta, C.; Biondo, C.; D'Andrea, E.; Dalmonech, D.; Marano, G.; Mäkelä, A.; Minunno, F.; et al. Accuracy, realism and general applicability of European forest models. *Glob. Chang. Biol.* **2022**, *28*, 6921–6943. [\[CrossRef\]](#) [\[PubMed\]](#)
43. Testolin, R.; Dalmonech, D.; Marano, G.; Bagnara, M.; D'Andrea, E.; Matteucci, G.; Noce, S.; Collalti, A. Simulating diverse forest management options in a changing climate on a *Pinus nigra* subsp. *laricio* plantation in Southern Italy. *Sci. Total Environ.* **2023**, *857*, 159361. [\[CrossRef\]](#) [\[PubMed\]](#)
44. Collalti, A.; Dalmonech, D.; Grieco, E.; Marano, G.; Vangi, E.; Puchi, P.; Orrico, M.R. 3D-CMCC-FEM(Coupled Model Carbon Cycle)BioGeoChemical and BiophysicalForest Ecosystem Model. 2023. Available online: <http://www.forest-modelling-lab.com> (accessed on 20 May 2024).
45. Farquhar, G.D.; von Caemmerer, S.; Berry, J.A. A biochemical model of photosynthetic CO<sub>2</sub> assimilation in leaves of C<sub>3</sub> species. *Planta* **1980**, *149*, 78–90. [\[CrossRef\]](#)
46. Bernacchi, C.J.; Singsaas, E.L.; Pimentel, C.; Portis, A.R.; Long, S.P. Improved temperature response functions for models of Rubisco-limited photosynthesis. *Plant. Cell Environ.* **2001**, *24*, 253–259. [\[CrossRef\]](#)
47. Bernacchi, C.J.; Calfapietra, C.; Davey, P.A.; Wittig, V.E.; Scarascia-Mugnozza, G.E.; Raines, C.A.; Long, S.P. Photosynthesis and stomatal conductance responses of poplars to free-air CO<sub>2</sub> enrichment (PopFACE) during the first growth cycle and immediately following coppice. *New Phytol.* **2003**, *159*, 609–621. [\[CrossRef\]](#) [\[PubMed\]](#)
48. Kattge, J.; Knorr, W. Temperature acclimation in a biochemical model of photosynthesis: A reanalysis of data from 36 species. *Plant. Cell Environ.* **2007**, *30*, 1176–1190. [\[CrossRef\]](#) [\[PubMed\]](#)
49. De Pury, D.G.G.; Farquhar, G.D. Simple scaling of photosynthesis from leaves to canopies without the errors of big-leaf models. *Plant. Cell Environ.* **1997**, *20*, 537–557. [\[CrossRef\]](#)
50. Amthor, J.S. The McCree-de Wit-Penning de Vries-Thornley respiration paradigms: 30 Years later. *Ann. Bot.* **2000**, *86*, 1–20. [\[CrossRef\]](#)
51. Smith, N.G.; Dukes, J.S. Plant respiration and photosynthesis in global-scale models: Incorporating acclimation to temperature and CO<sub>2</sub>. *Glob. Chang. Biol.* **2013**, *19*, 45–63. [\[CrossRef\]](#) [\[PubMed\]](#)
52. Merganičová, K.; Merganič, J.; Lehtonen, A.; Vacchiano, G.; Sever, M.Z.O.; Augustynczyk, A.L.D.; Grote, R.; Kyselová, I.; Mäkelä, A.; Yousefpour, R.; et al. Forest carbon allocation modelling under climate change. *Tree Physiol.* **2019**, *39*, 1937–1960. [\[CrossRef\]](#) [\[PubMed\]](#)
53. Parton, W.J.; Scurlock, J.M.O.; Ojima, D.S.; Gilmanov, T.G.; Scholes, R.J.; Schimel, D.S.; Kirchner, T.; Menaut, J.-C.; Seastedt, T.; Garcia Moya, E.; et al. Observations and modeling of biomass and soil organic matter dynamics for the grassland biome worldwide. *Glob. Biogeochem. Cycles* **1993**, *7*, 785–809. [\[CrossRef\]](#)
54. Parton, W.J.; Schimel, D.S.; Cole, C.V.; Ojima, D.S. Analysis of Factors Controlling Soil Organic Matter Levels in Great Plains Grasslands. *Soil Sci. Soc. Am. J.* **1987**, *51*, 1173–1179. [\[CrossRef\]](#)
55. Thornton, P.E.; Rosenbloom, N.A. Ecosystem model spin-up: Estimating steady state conditions in a coupled terrestrial carbon and nitrogen cycle model. *Ecol. Modell.* **2005**, *189*, 25–48. [\[CrossRef\]](#)
56. Pastorello, G.; Trotta, C.; Canfora, E.; Chu, H.; Christianson, D.; Cheah, Y.W.; Poindexter, C.; Chen, J.; Elbashandy, A.; Humphrey, M.; et al. The FLUXNET2015 dataset and the ONEFlux processing pipeline for eddy covariance data. *Sci. Data* **2020**, *7*, 225. [\[CrossRef\]](#)
57. Frieler, K.; Lange, S.; Piontek, F.; Reyer, C.P.O.; Schewe, J.; Warszawski, L.; Zhao, F.; Chini, L.; Denvil, S.; Emanuel, K.; et al. Assessing the impacts of 1.5 °C global warming—Simulation protocol of the Inter-Sectoral Impact Model Intercomparison Project (ISIMIP2b). *Geosci. Model Dev.* **2017**, *10*, 4321–4345. [\[CrossRef\]](#)
58. Reyer, C.P.O.; Silveyra Gonzalez, R.; Dolos, K.; Hartig, F.; Hauf, Y.; Noack, M.; Lasch-Born, P.; Rötzer, T.; Pretzsch, H.; Meesenburg, H.; et al. The PROFOUND Database for evaluating vegetation models and simulating climate impacts on European forests. *Earth Syst. Sci. Data* **2020**, *12*, 1295–1320. [\[CrossRef\]](#)

59. Grünig, M.; Rammer, W.; Albrich, K.; André, F.; Augustynczyk, A.L.D.; Bohn, F.; Bouwman, M.; Bugmann, H.; Collalti, A.; Cristal, I.; et al. A harmonized database of European forest simulations under climate change. *Data Br.* **2024**, *54*, 110384. [\[CrossRef\]](#)
60. ESA. Forest Type 2018 (Raster 10 m), Europe, 3-Yearly, Oct. 2020. Available online: <https://sdi.eea.europa.eu/catalogue/copernicus/api/records/db1af59f-f01f-4bd4-830c-f0eb652500c1?language=all> (accessed on 21 March 2024).
61. Moss, R.H.; Edmonds, J.A.; Hibbard, K.A.; Manning, M.R.; Rose, S.K.; Van Vuuren, D.P.; Carter, T.R.; Emori, S.; Kainuma, M.; Kram, T.; et al. The next generation of scenarios for climate change research and assessment. *Nature* **2010**, *463*, 747–756. [\[CrossRef\]](#) [\[PubMed\]](#)
62. van Vuuren, D.P.; Edmonds, J.; Kainuma, M.; Riahi, K.; Thomson, A.; Hibbard, K.; Hurtt, G.C.; Kram, T.; Krey, V.; Lamarque, J.F.; et al. The representative concentration pathways: An overview. *Clim. Change* **2011**, *109*, 5–31. [\[CrossRef\]](#)
63. Hempel, S.; Frieler, K.; Warszawski, L.; Schewe, J.; Piontek, F. A trend-preserving bias correction—The ISI-MIP approach. *Earth Syst. Dyn.* **2013**, *4*, 219–236. [\[CrossRef\]](#)
64. Lange, S. Bias correction of surface downwelling longwave and shortwave radiation for the EWEMBI dataset. *Earth Syst. Dyn.* **2018**, *9*, 627–645. [\[CrossRef\]](#)
65. Meinshausen, M.; Smith, S.J.; Calvin, K.; Daniel, J.S.; Kainuma, M.L.T.; Lamarque, J.; Matsumoto, K.; Montzka, S.A.; Raper, S.C.B.; Riahi, K.; et al. The RCP greenhouse gas concentrations and their extensions from 1765 to 2300. *Clim. Change* **2011**, *109*, 213–241. [\[CrossRef\]](#)
66. Dlugokencky, E.; Tans, P. *Trends in Atmospheric Carbon Dioxide*; National Oceanic & Atmospheric Administration, Earth System Research Laboratory (NOAA/ESRL): Washington, DC, USA, 2014.
67. Reichstein, M.; Falge, E.; Baldocchi, D.; Papale, D.; Aubinet, M.; Berbigier, P.; Bernhofer, C.; Buchmann, N.; Gilmanov, T.; Granier, A.; et al. On the separation of net ecosystem exchange into assimilation and ecosystem respiration: Review and improved algorithm. *Glob. Chang. Biol.* **2005**, *11*, 1424–1439. [\[CrossRef\]](#)
68. Papale, D.; Reichstein, M.; Aubinet, M.; Canfora, E.; Bernhofer, C.; Kutsch, W.; Longdoz, B.; Rambal, S.; Valentini, R.; Vesala, T.; et al. Towards a standardized processing of Net Ecosystem Exchange measured with eddy covariance technique: Algorithms and uncertainty estimation. *Biogeosciences* **2006**, *3*, 571–583. [\[CrossRef\]](#)
69. Nolè, A.; Collalti, A.; Magnani, F.; Duce, P.; Ferrara, A.; Mancino, G.; Marras, S.; Sirca, C.; Spano, D.; Borghetti, M. Assessing temporal variation of primary and ecosystem production in two Mediterranean forests using a modified 3-PG model. *Ann. For. Sci.* **2013**, *70*, 729–741. [\[CrossRef\]](#)
70. Jiang, F.; Ju, W.; He, W.; Wu, M.; Wang, H.; Wang, J.; Jia, M.; Feng, S.; Zhang, L.; Chen, J.M. A 10-year global monthly averaged terrestrial net ecosystem exchange dataset inferred from the ACOS GOSAT v9 XCO<sub>2</sub> retrievals (GCAS2021). *Earth Syst. Sci. Data* **2022**, *14*, 3013–3037. [\[CrossRef\]](#)
71. Jarvis, P.; Rey, A.; Petsikos, C.; Wingate, L.; Rayment, M.; Pereira, J.; Banza, J.; David, J.; Miglietta, F.; Borghetti, M.; et al. Drying and wetting of Mediterranean soils stimulates decomposition and carbon dioxide emission: The “Birch effect”. *Tree Physiol.* **2007**, *27*, 929–940. [\[CrossRef\]](#) [\[PubMed\]](#)
72. Lavigne, M.B.; Ryan, M.G.; Anderson, D.E.; Baldocchi, D.D.; Crill, P.M.; Fitzjarrald, D.R.; Goulden, M.L.; Gower, S.T.; Massheder, J.M.; McCaughey, J.H.; et al. Comparing nocturnal eddy covariance measurements to estimates of ecosystem respiration made by scaling chamber measurements at six coniferous boreal sites. *J. Geophys. Res. Atmos.* **1997**, *102*, 28977–28985. [\[CrossRef\]](#)
73. Speckman, H.N.; Frank, J.M.; Bradford, J.B.; Miles, B.L.; Massman, W.J.; Parton, W.J.; Ryan, M.G. Forest ecosystem respiration estimated from eddy covariance and chamber measurements under high turbulence and substantial tree mortality from bark beetles. *Glob. Chang. Biol.* **2015**, *21*, 708–721. [\[CrossRef\]](#) [\[PubMed\]](#)
74. Campioli, M.; Malhi, Y.; Vicca, S.; Luyssaert, S.; Papale, D.; Peñuelas, J.; Reichstein, M.; Migliavacca, M.; Arain, M.A.; Janssens, I.A. Evaluating the convergence between eddy-covariance and biometric methods for assessing carbon budgets of forests. *Nat. Commun.* **2016**, *7*, 13717. [\[CrossRef\]](#)
75. Ryan, M.G. The enduring mystery of differences between eddy covariance and biometric measurements for ecosystem respiration and net carbon storage in forests. *New Phytol.* **2023**, *239*, 2060–2063. [\[CrossRef\]](#)
76. Cheng, W.; Huang, L.; Liu, Z.; Dong, J.; Moore, J.C.; MacMartin, D.G.; Deng, X. Seasonal and regional changes in terrestrial carbon uptake under an overshoot scenario. *Resour. Conserv. Recycl.* **2023**, *195*, 106997. [\[CrossRef\]](#)
77. van der Woude, A.M.; Peters, W.; Joetzer, E.; Lafont, S.; Koren, G.; Ciais, P.; Ramonet, M.; Xu, Y.; Bastos, A.; Botía, S.; et al. Temperature extremes of 2022 reduced carbon uptake by forests in Europe. *Nat. Commun.* **2023**, *14*, 6218. [\[CrossRef\]](#)
78. Wolf, S.; Paul-Limoges, E. Drought and heat reduce forest carbon uptake. *Nat. Commun.* **2023**, *14*, 14–17. [\[CrossRef\]](#) [\[PubMed\]](#)
79. Chen, Z.; Wang, W.; Forzieri, G.; Cescatti, A. Transition from positive to negative indirect CO<sub>2</sub> effects on the vegetation carbon uptake. *Nat. Commun.* **2024**, *15*, 1500. [\[CrossRef\]](#) [\[PubMed\]](#)
80. De Marco, A.; Sicard, P.; Feng, Z.; Agathokleous, E.; Alonso, R.; Araminiene, V.; Augustatis, A.; Badea, O.; Beasley, J.C.; Branquinho, C.; et al. Strategic roadmap to assess forest vulnerability under air pollution and climate change. *Glob. Chang. Biol.* **2022**, *28*, 5062–5085. [\[CrossRef\]](#) [\[PubMed\]](#)
81. Nissan, A.; Alcolombri, U.; Peleg, N.; Galili, N.; Jimenez-Martinez, J.; Molnar, P.; Holzner, M. Global warming accelerates soil heterotrophic respiration. *Nat. Commun.* **2023**, *14*, 3452. [\[CrossRef\]](#) [\[PubMed\]](#)
82. Ping, J.; Cui, E.; Du, Y.; Wei, N.; Zhou, J.; Wang, J.; Niu, S.; Luo, Y.; Xia, J. Enhanced causal effect of ecosystem photosynthesis on respiration during heatwaves. *Sci. Adv.* **2023**, *9*, eadi6395. [\[CrossRef\]](#) [\[PubMed\]](#)
83. Odum, E.P. The strategy of ecosystem development. *Science* **1969**, *164*, 262–270. [\[CrossRef\]](#) [\[PubMed\]](#)

84. McGrath, M.J.; Schulte-Frohlinde, A.; Luyssaert, S. New ways for (in)validating the forest carbon neutrality hypothesis. *Glob. Chang. Biol.* **2024**, *30*, e16982. [[CrossRef](#)] [[PubMed](#)]
85. Gundersen, P.; Thybring, E.E.; Nord-Larsen, T.; Vesterdal, L.; Nadelhoffer, K.J.; Johannsen, V.K. *Old-Growth Forest Carbon Sinks Overestimated*; Nature Publishing Group: New York, NY, USA, 2021; Volume 591, pp. E21–E23.
86. Kirschbaum, M.U.F. Direct and indirect climate change effects on photosynthesis and transpiration. *Plant Biol.* **2004**, *6*, 242–253. [[CrossRef](#)] [[PubMed](#)]
87. Garrity, S.R.; Bohrer, G.; Maurer, K.D.; Mueller, K.L.; Vogel, C.S.; Curtis, P.S. A comparison of multiple phenology data sources for estimating seasonal transitions in deciduous forest carbon exchange. *Agric. For. Meteorol.* **2011**, *151*, 1741–1752. [[CrossRef](#)]
88. Weir, J.C.; Phillimore, A.B. Buffering and phenological mismatch: A change of perspective. *Glob. Chang. Biol.* **2024**, *30*, e17294. [[CrossRef](#)]
89. Bowling, D.R.; Schädel, C.; Smith, K.R.; Richardson, A.D.; Bahn, M.; Arain, M.A.; Varlagin, A.; Ouimette, A.P.; Frank, J.M.; Barr, A.G.; et al. Phenology of Photosynthesis in Winter-Dormant Temperate and Boreal Forests: Long-Term Observations from Flux Towers and Quantitative Evaluation of Phenology Models. *J. Geophys. Res. Biogeosci.* **2024**, *129*, e2023JG007839. [[CrossRef](#)]
90. Beamesderfer, E.R.; Arain, M.A.; Khomik, M.; Brodeur, J.J. The Impact of Seasonal and Annual Climate Variations on the Carbon Uptake Capacity of a Deciduous Forest Within the Great Lakes Region of Canada. *J. Geophys. Res. Biogeosci.* **2020**, *125*, e2019JG005389. [[CrossRef](#)] [[PubMed](#)]
91. Luo, X.; Zhao, R.; Chu, H.; Fatichi, S.; Keenan, T.; Lu, X.; Nguyen, N.; Prentice, I.; Yu, L. Deciduous forests use carbon more efficiently than evergreen forests. *Preprint* **2024**. [[CrossRef](#)]
92. Terrer, C.; Vicca, S.; Hungate, B.A.; Phillips, R.P.; Prentice, I.C. Mycorrhizal association as a primary control of the CO<sub>2</sub> fertilization effect. *Science* **2016**, *353*, 72–74. [[CrossRef](#)]

**Disclaimer/Publisher’s Note:** The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.